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Nitrogen input to croplands and its environmental impacts in China



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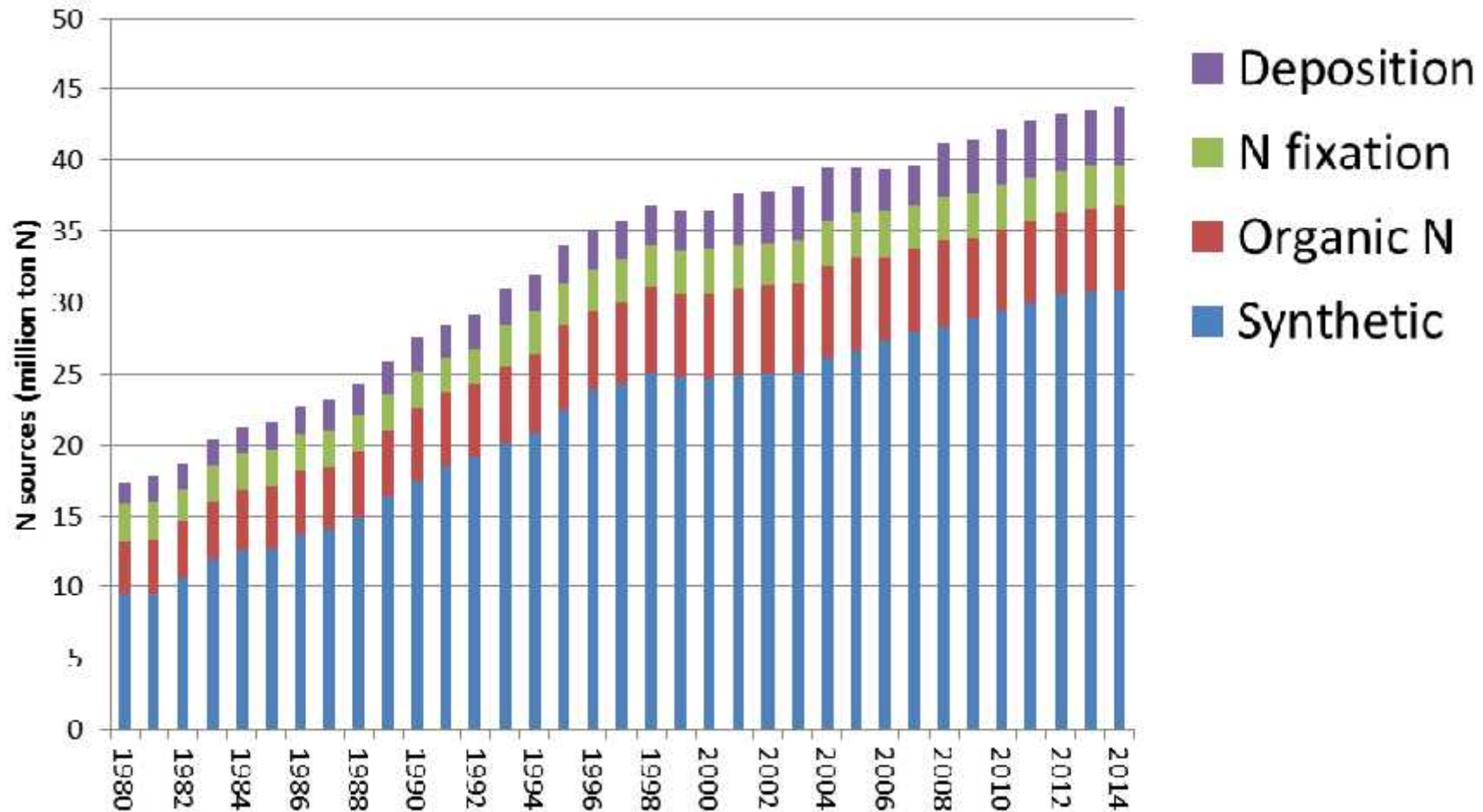
Outlines

- Spatial and temporal changes in N input and NUE
- Efforts in improving NUE
- Environmental impacts
 - Biodiversity loss
 - Soil acidification
 - Greenhouse gas balance
 - Eutrophication
 - Air pollution

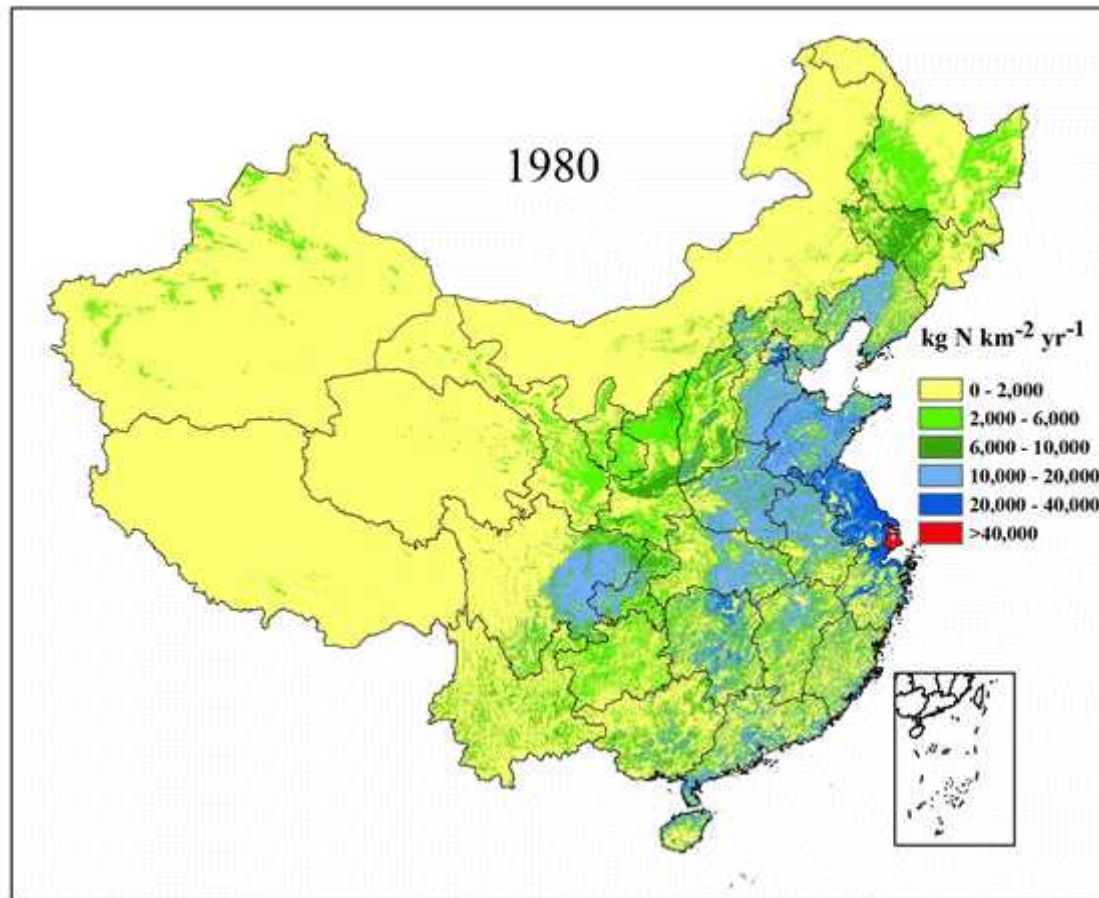
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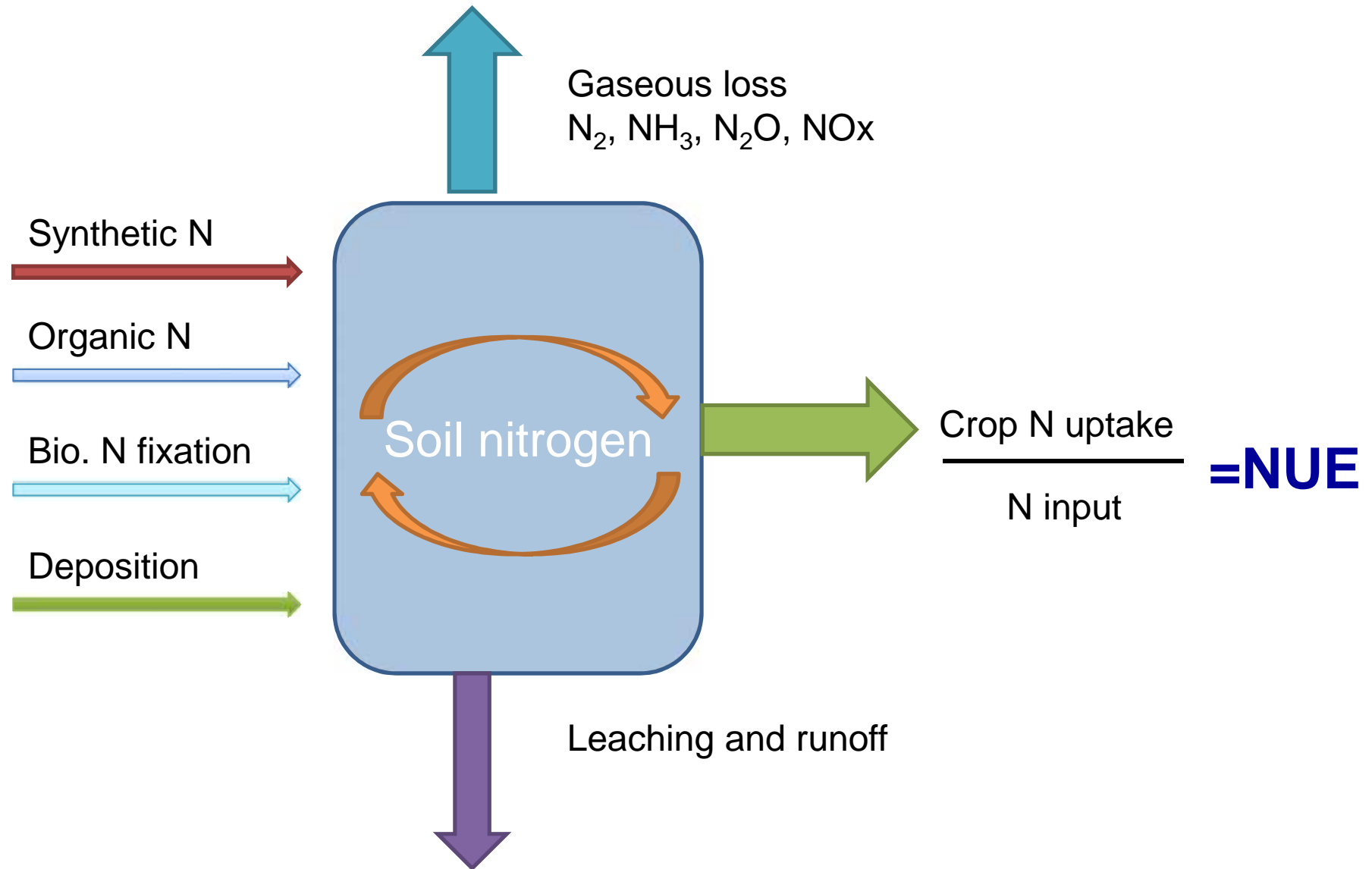
Nitrogen input to croplands



Spatial and temporal variation of N input to croplands



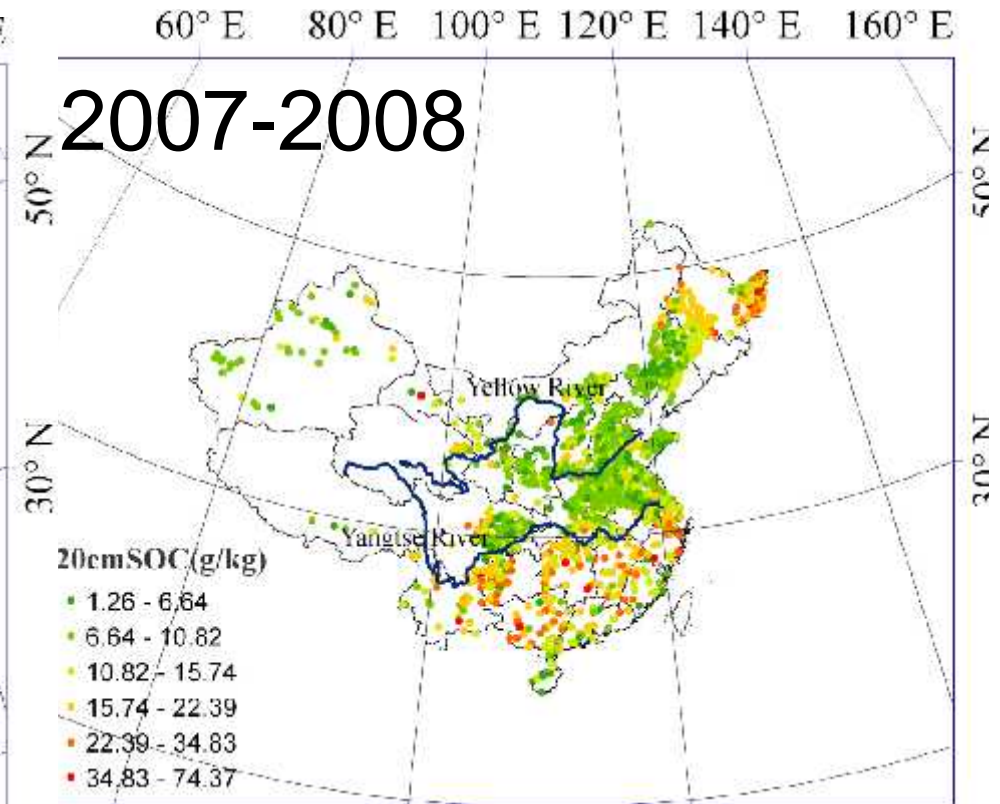
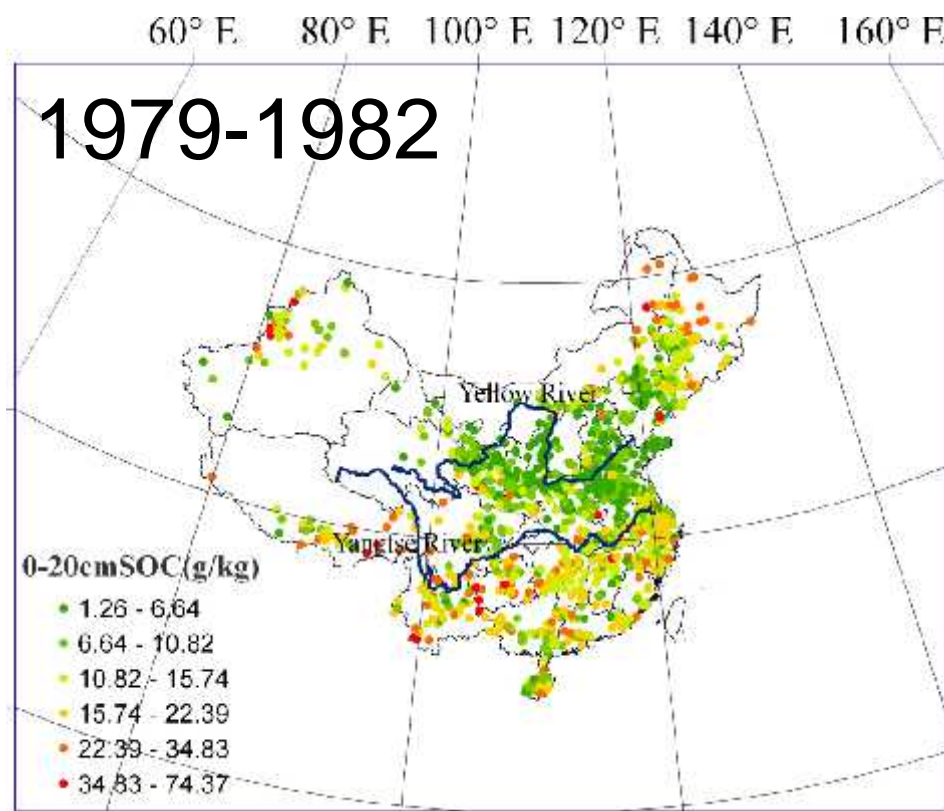
Nitrogen use efficiency (NUE)



SOC and soil N content changes

National soil survey

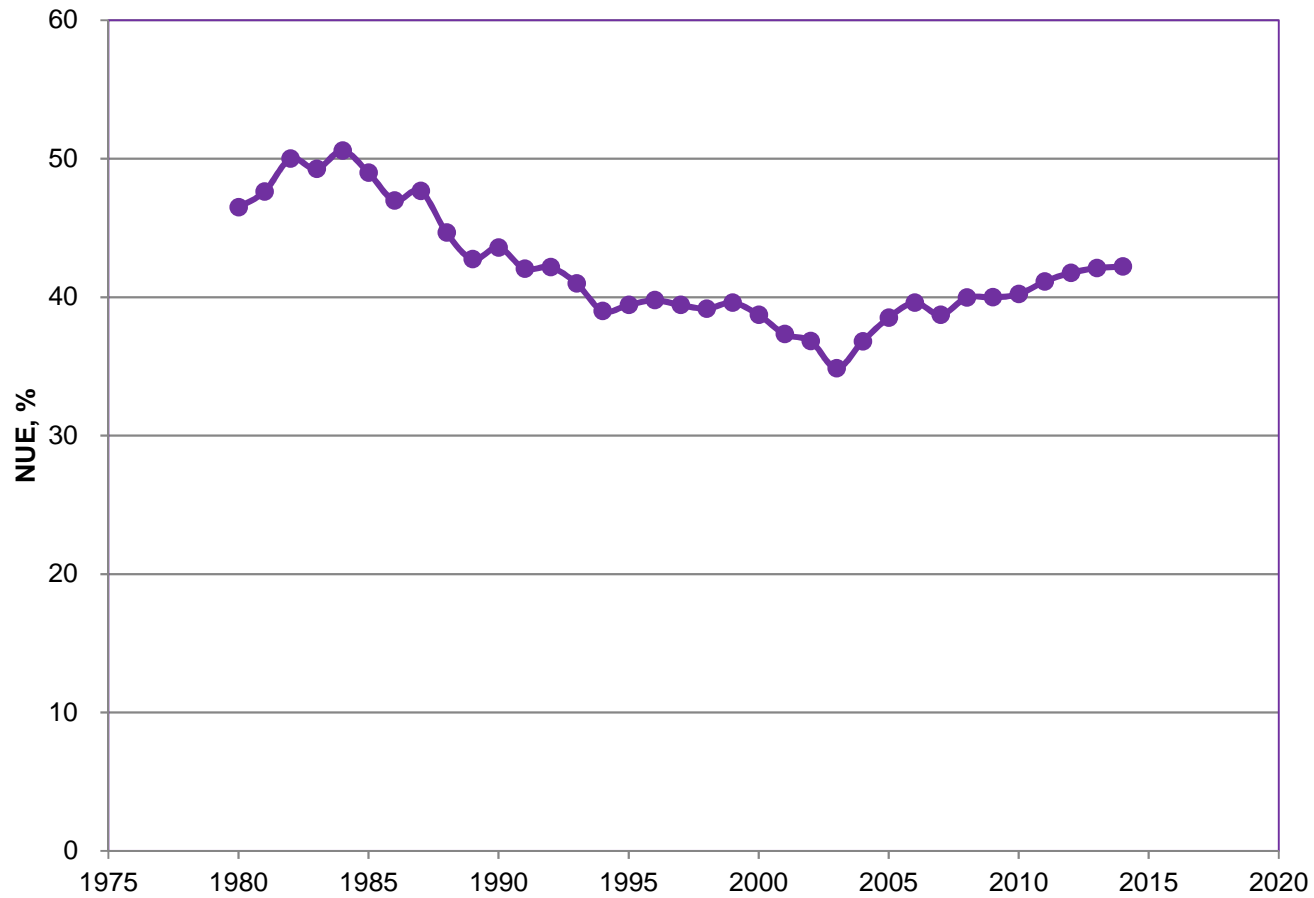
Soil sampling campaign



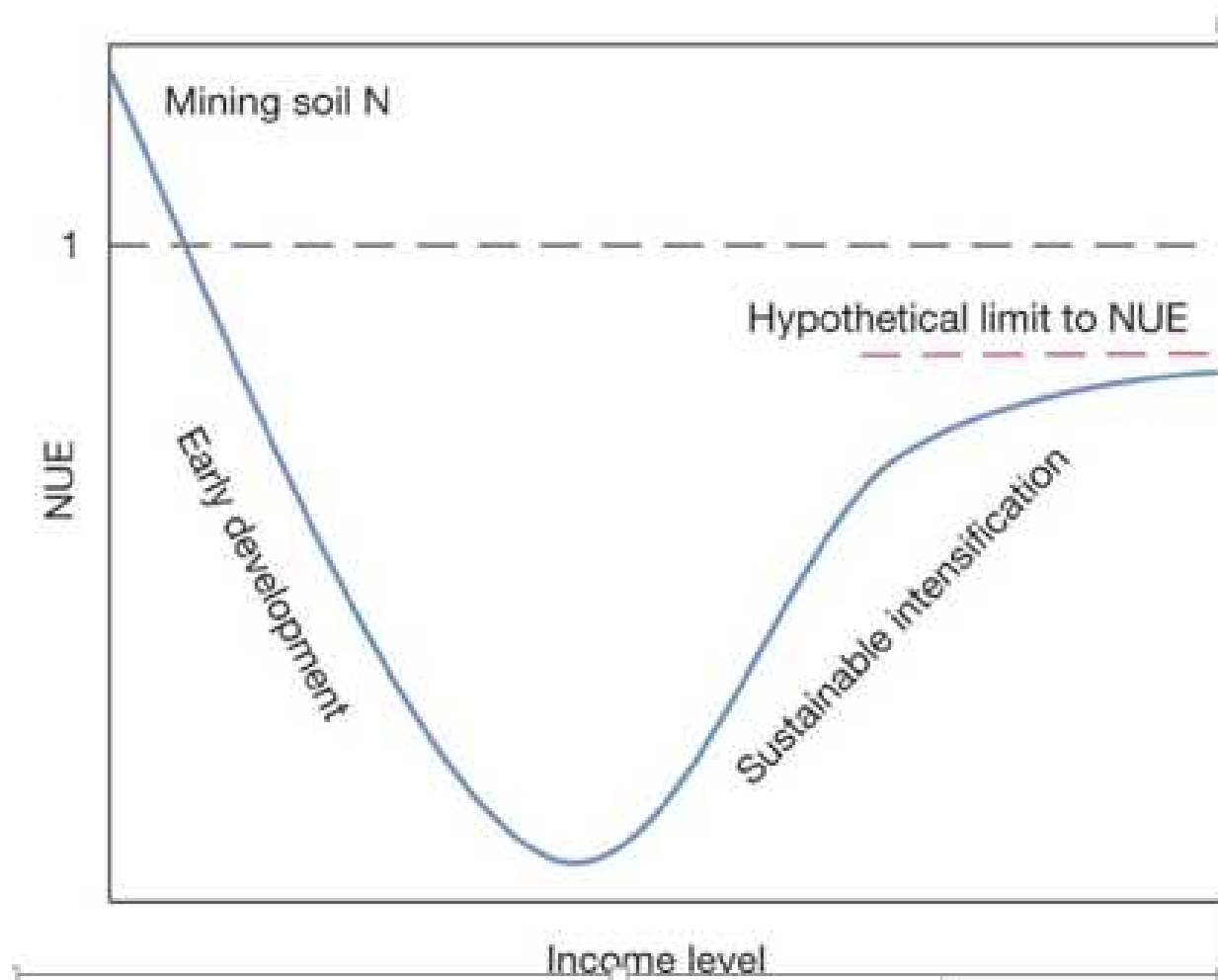
C:+8.8%

N:+5.1%

Change in NUE with time

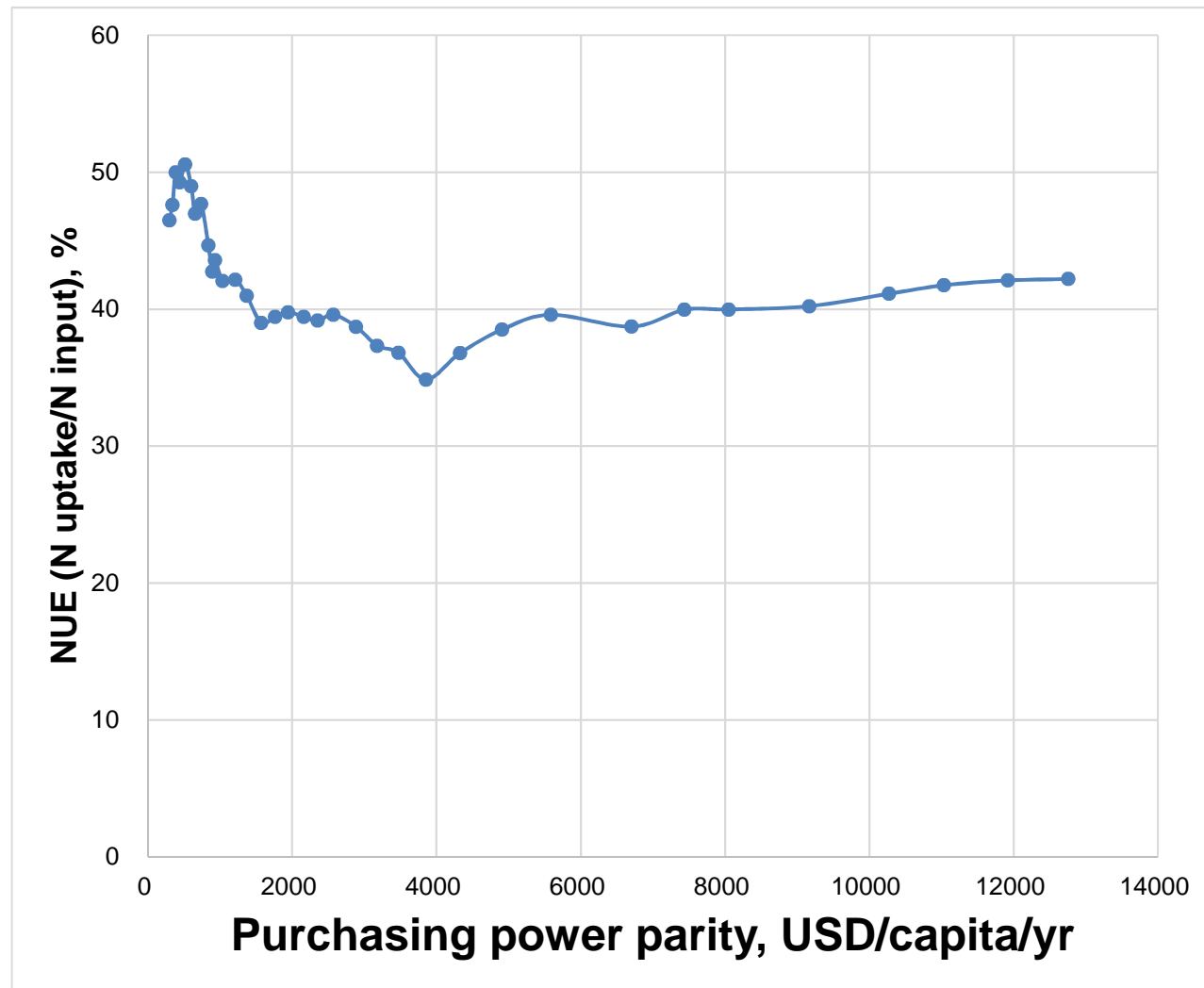


Environmental Kuznets Curve



Zhang et al., 2015, Nature 528, 51–59

Changes in NUE with economy growth

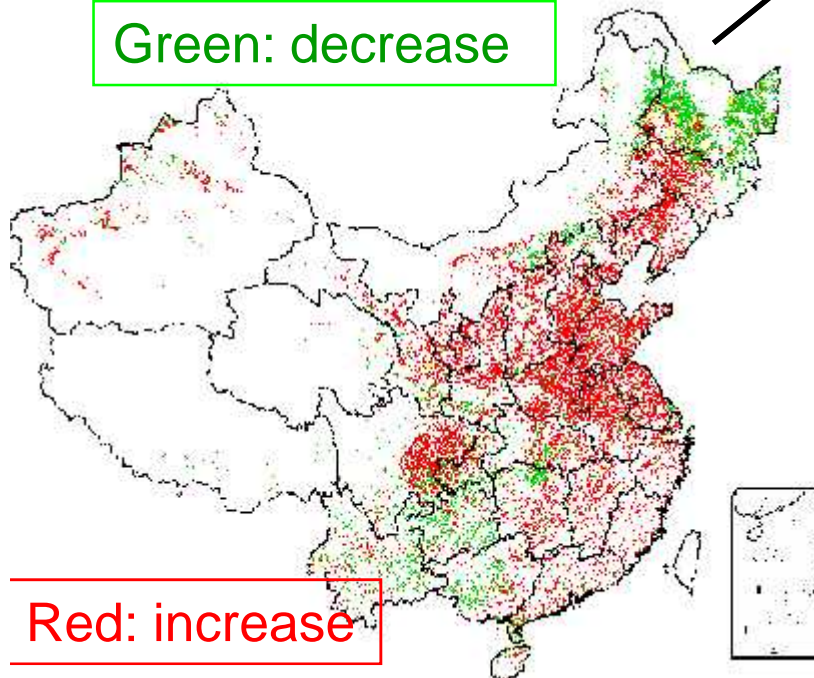


Heilongjiang and Inner Mongolia

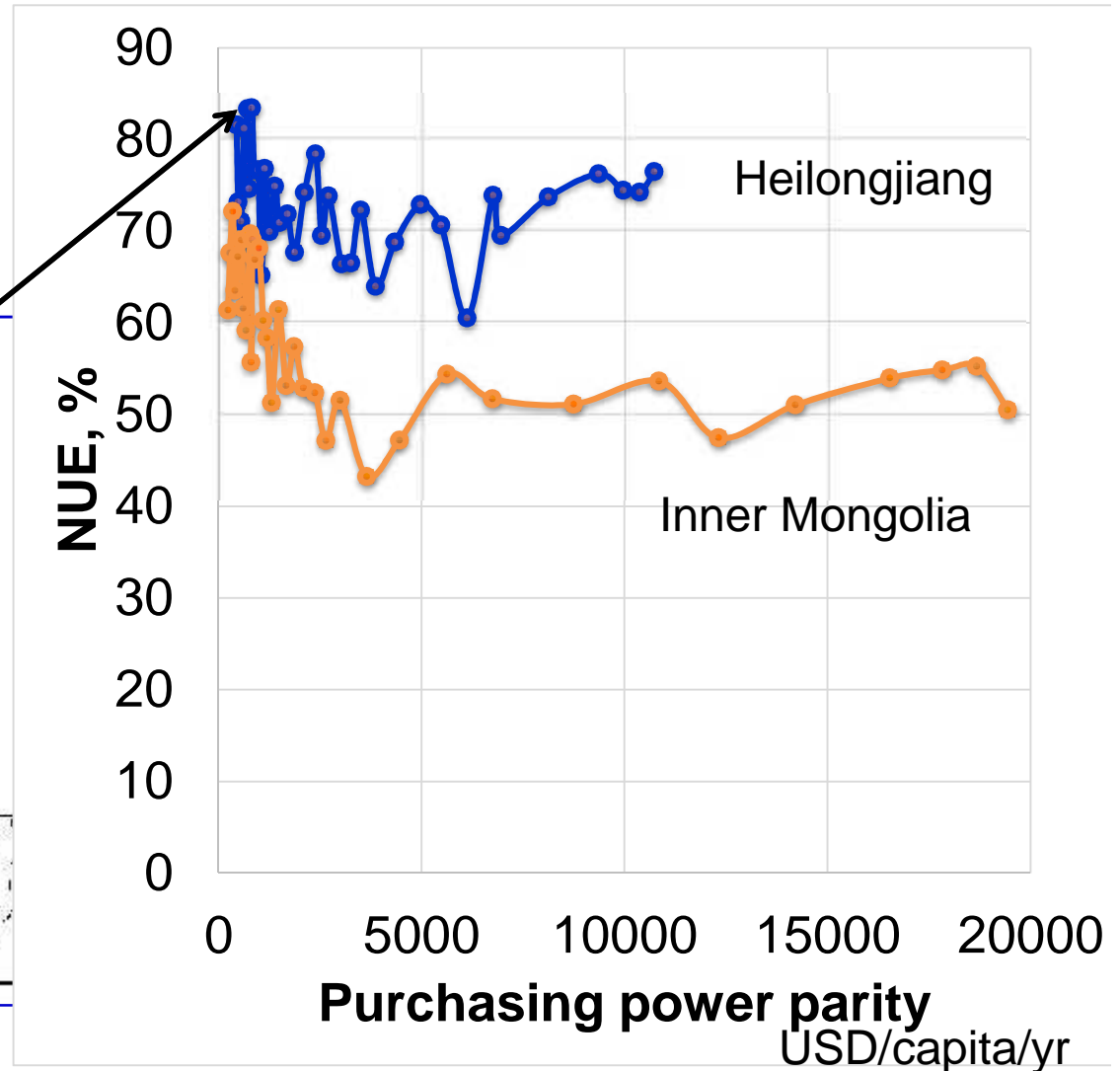
Consistent high NUE
Loss of soil nitrogen

Green: decrease

Red: increase

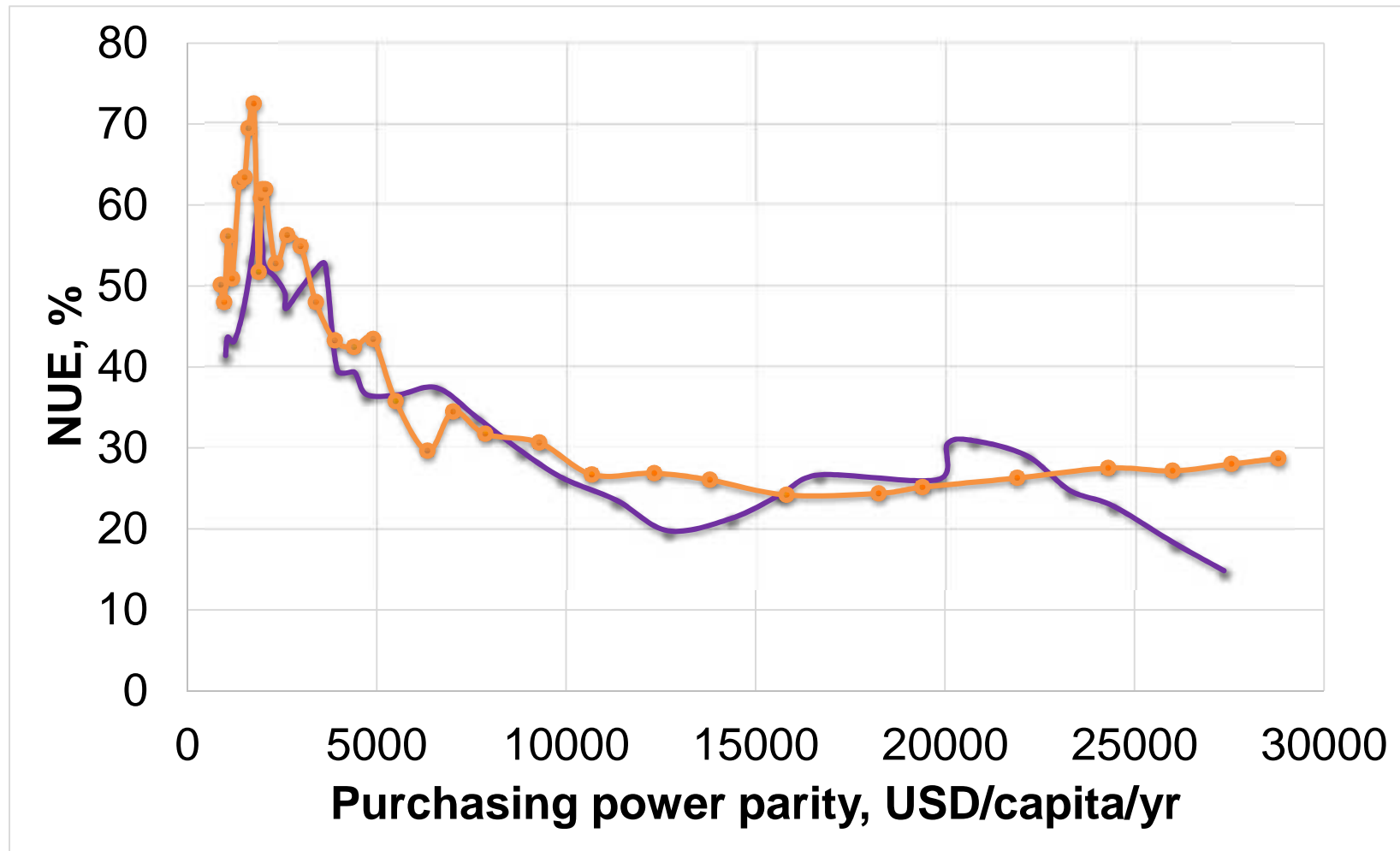


Courtesy of Huang Yao



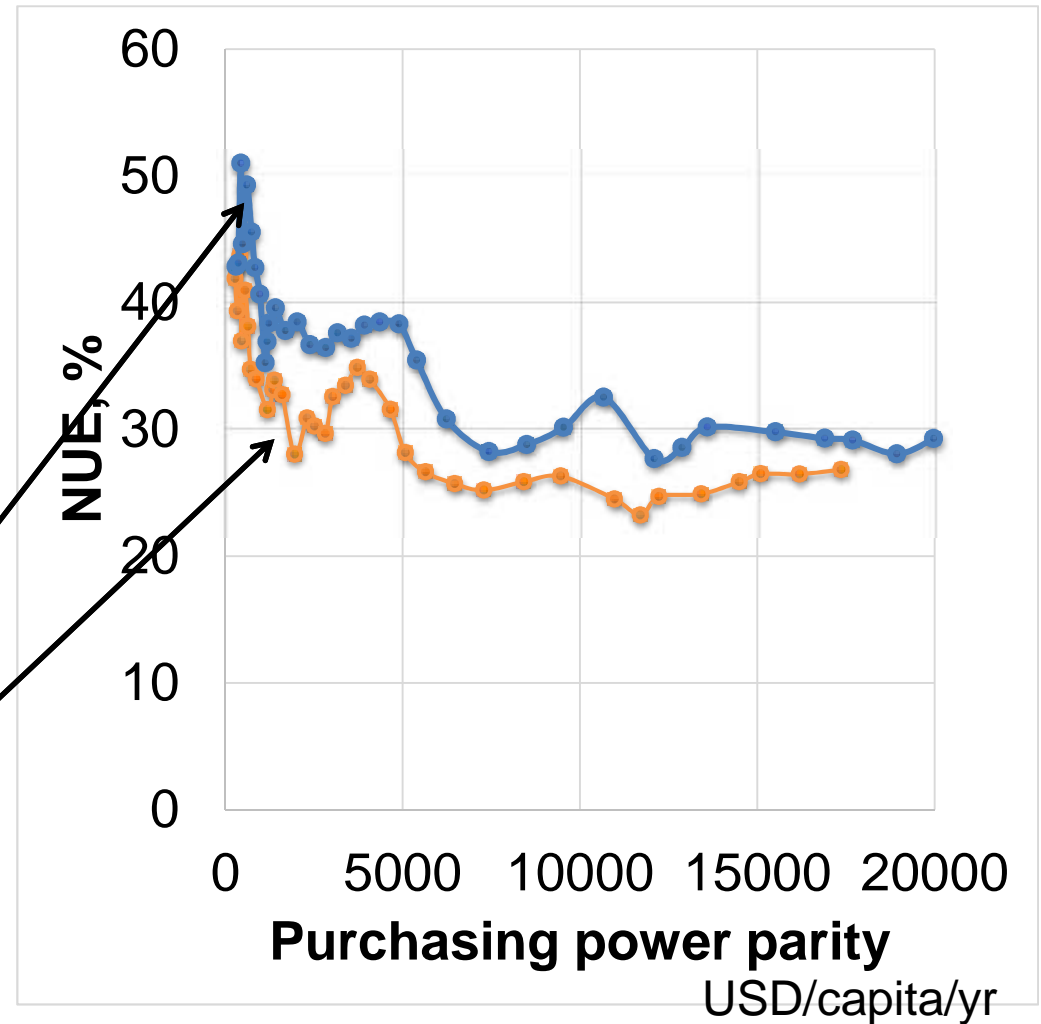
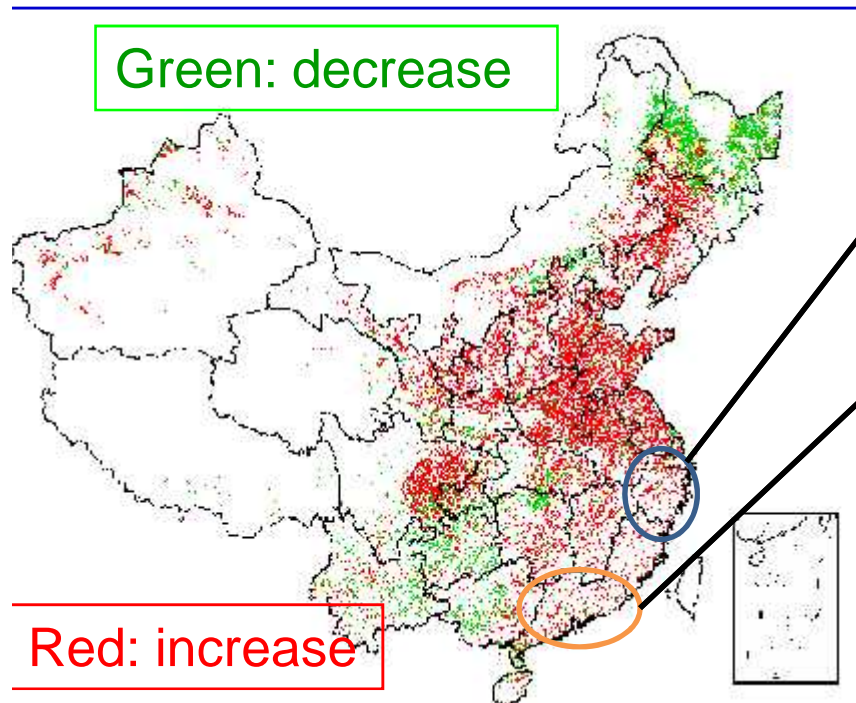
Shanghai and Beijing

Urban area, economically developed, consistent low NUE



Zhejiang and Guangdong

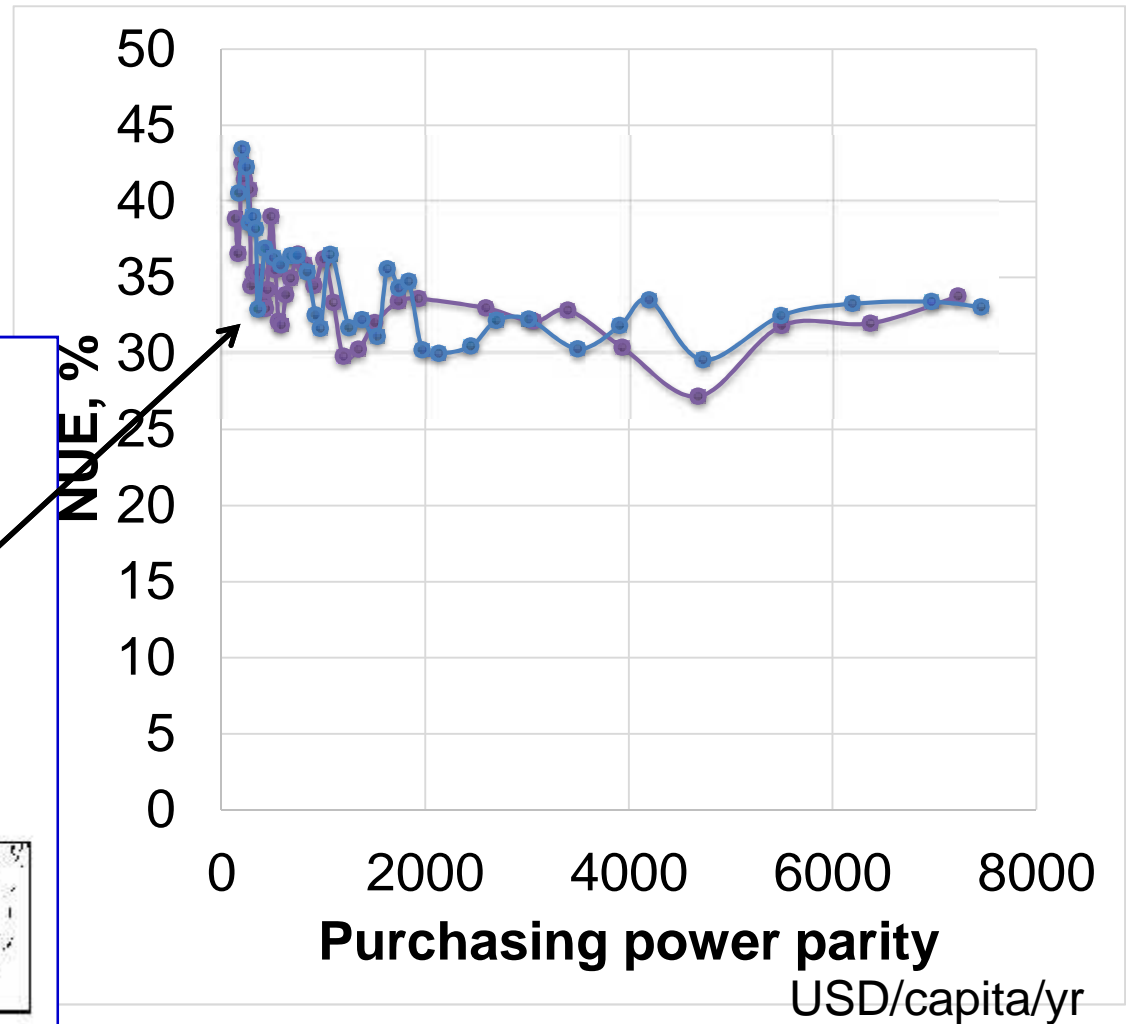
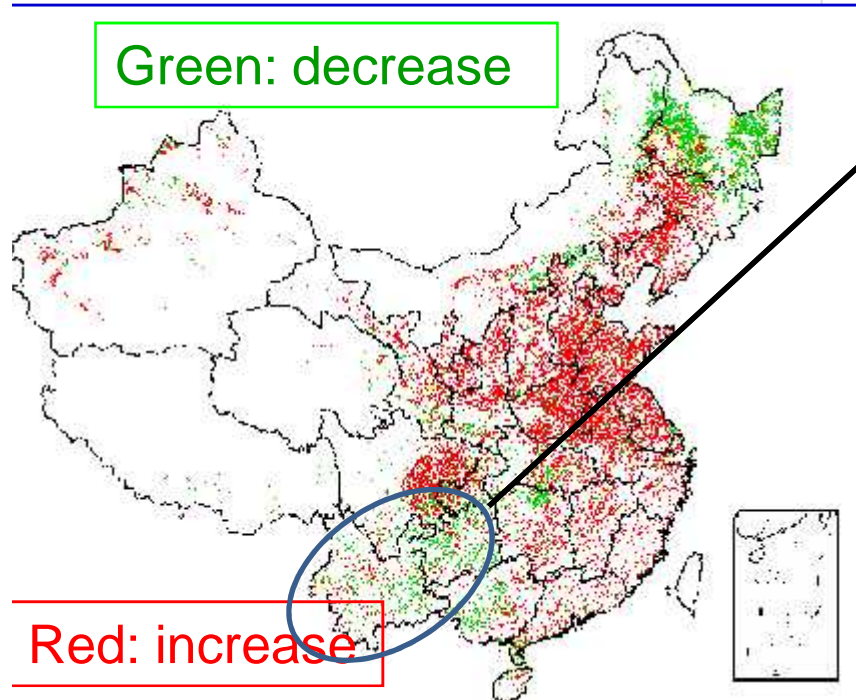
Economically developed
Soil N little change
Low NUE
Not increasing yet



Courtesy of Huang Yao

Yunnan and Guizhou

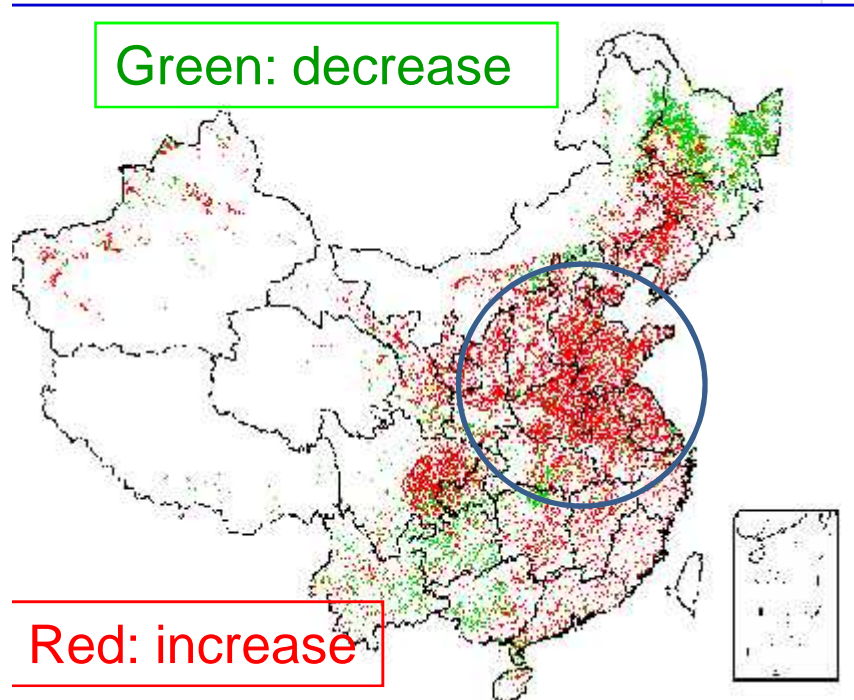
Economically undeveloped
Slight decrease in soil N
Low and stable NUE



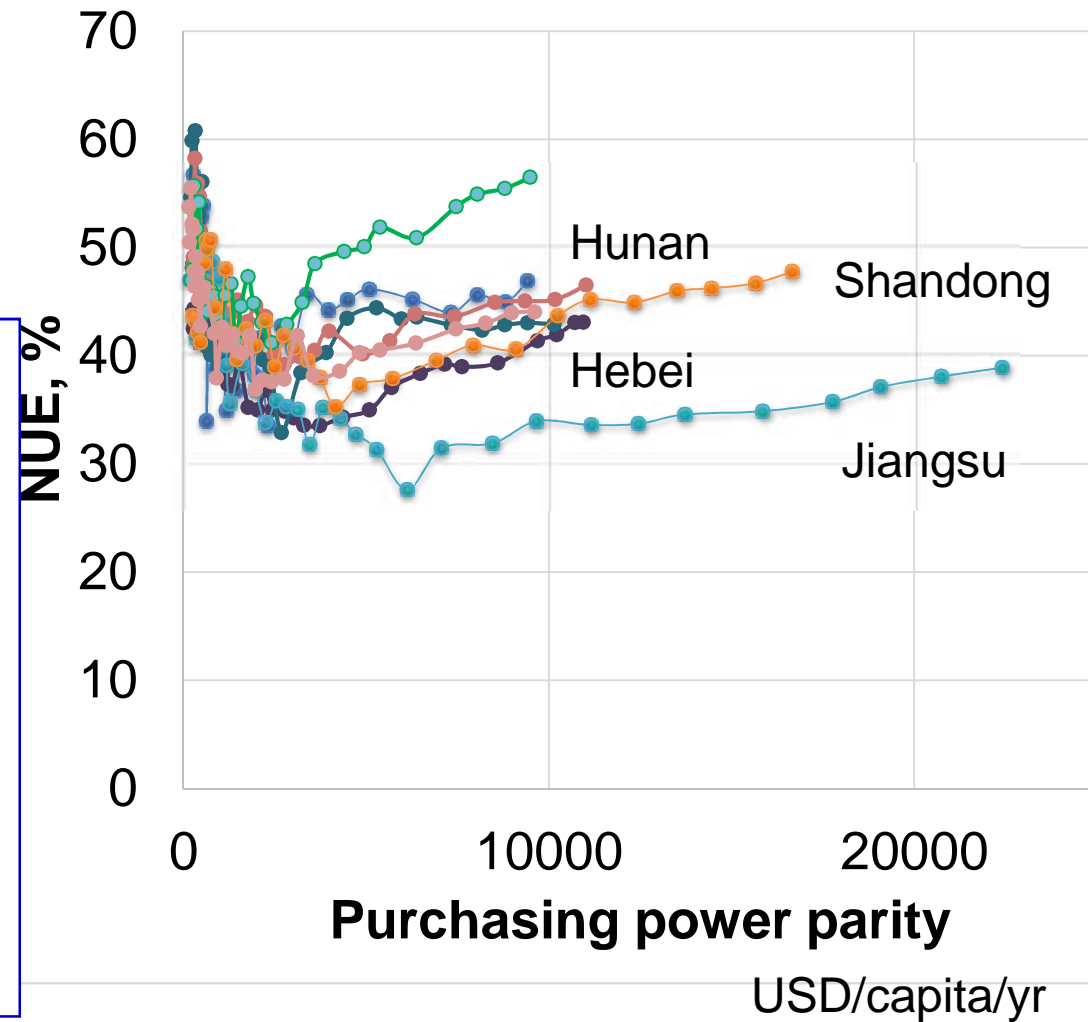
Courtesy of Huang Yao

Major agricultural region

High or middle level economy
Major agricultural provinces
Largely increased soil N
Increasing NUE



Courtesy of Huang Yao



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Knowledge-based N practices

- **Enhanced efficiency N fertilizers**

 - Controlled release N fertilizer

 - Nitrification inhibitor

 - Urease inhibitor

- **Optimized N application**

 - Reducing basal fertilizer N ratio

 - Increasing N splitting frequency

 - Deep placement of fertilizer

 - Fertilizer recommendation based on soil test

Fertilization based on soil test program



An accumulative area of 100 million hector by 2015

Fertilizer recommendation systems



Computer access



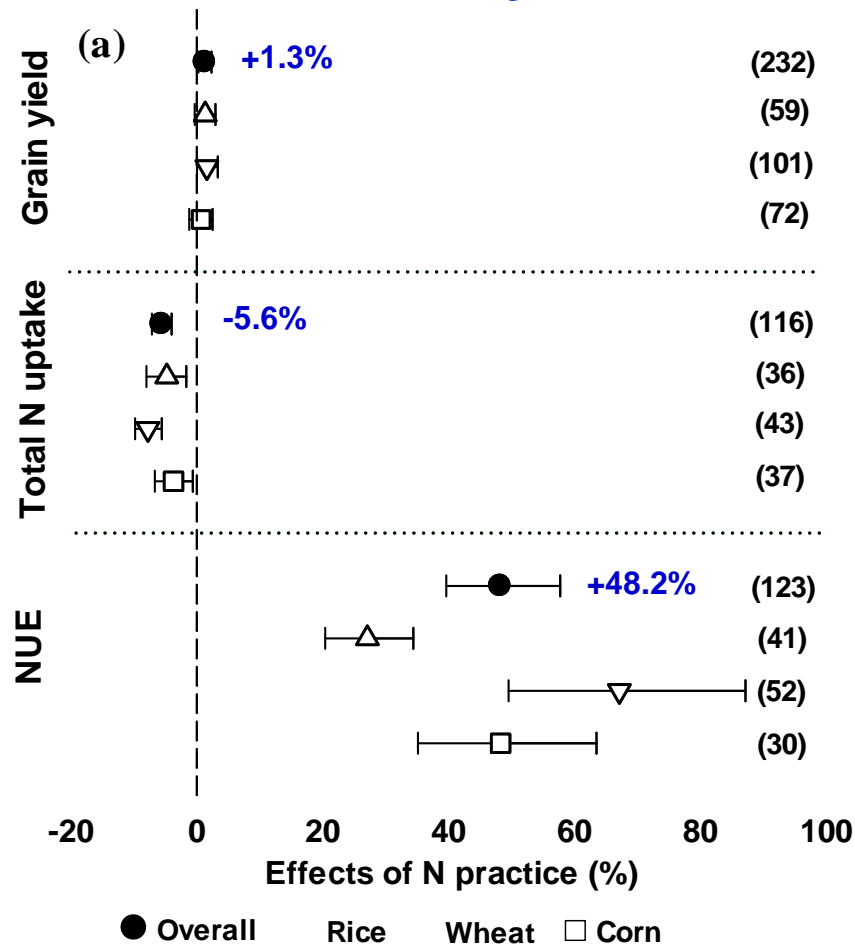
Palm access

Touching panel

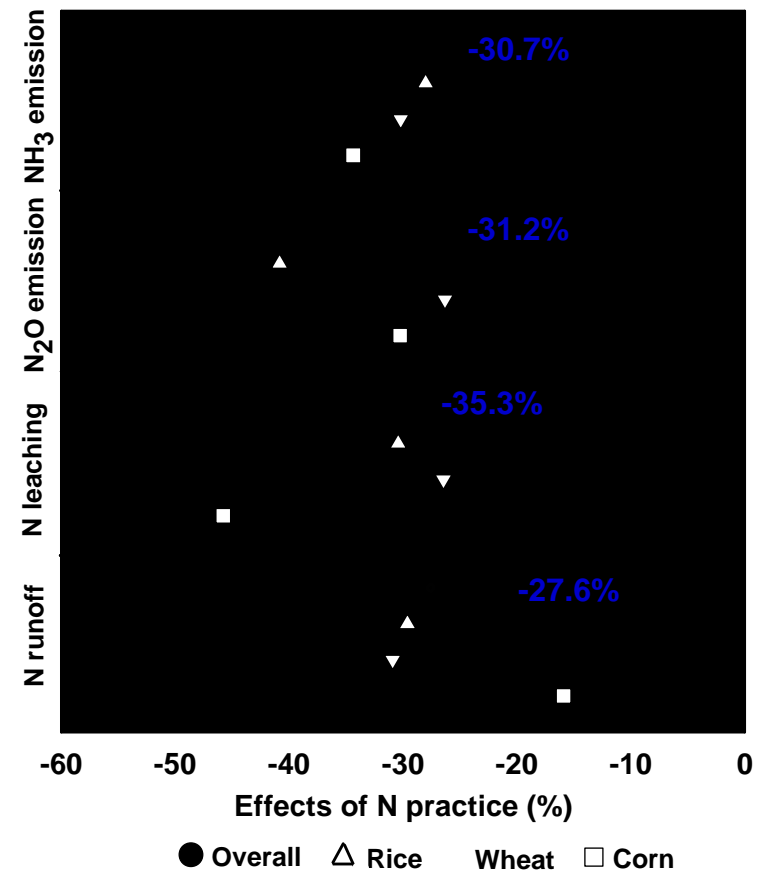


Effects on productivity and N loss

Productivity indexes



Loss indexes



Achieved on an average N reduction of 28%

Barriers

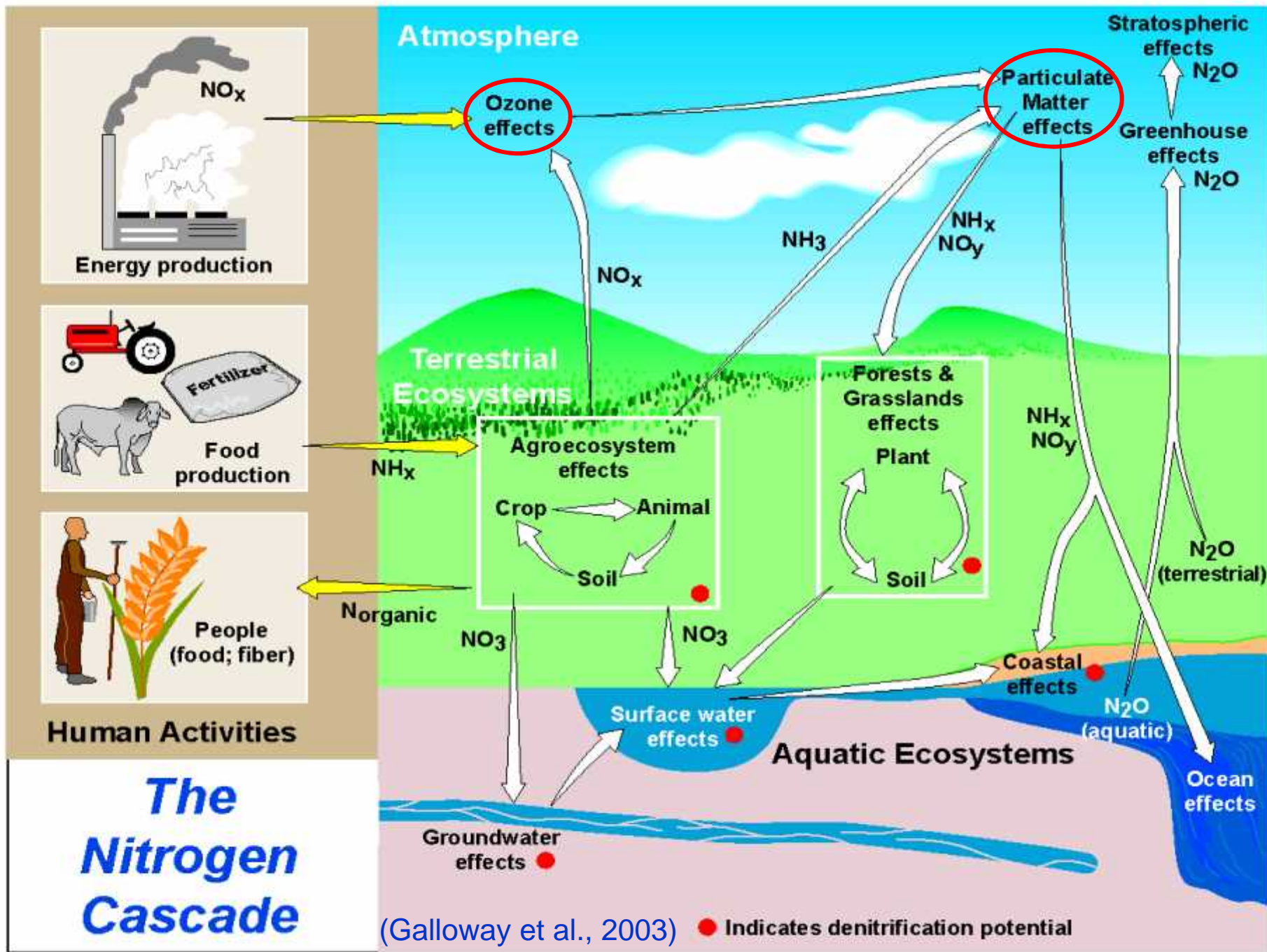
- **Effects of these N practices varied among different crop species and soil properties.**
- **Many farms are still small scale, farmers' knowledge, environmental awareness still need to improve**
- **Opportunity cost (time, labor, training/education costs) for implementing**
- **Lack of mechanization**



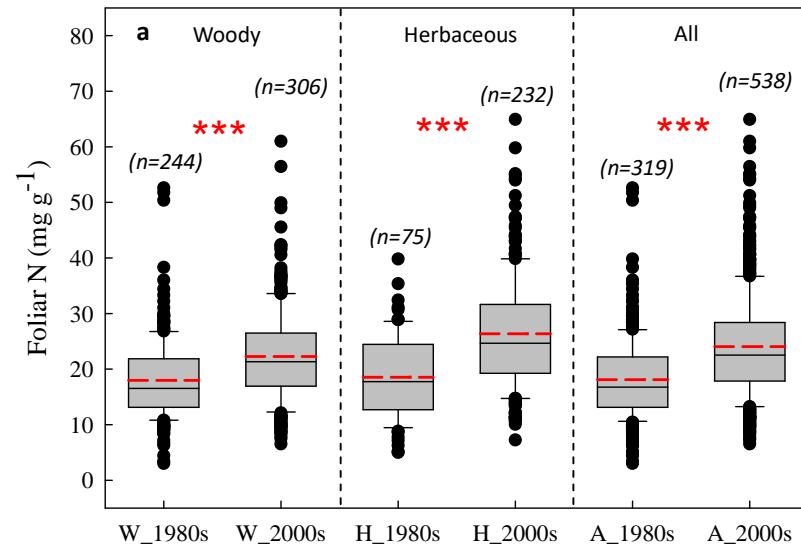
National geography, 2011

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Significant increase in leaf N in natural vegetation (a) and N uptake in long-term unfertilized croplands (b) during 1980s and 2000s.



Foliar N concentration

1980s: $18.1 \pm 7.2 \text{ mg/g}$

2000s: $24.0 \pm 9.2 \text{ mg/g}$

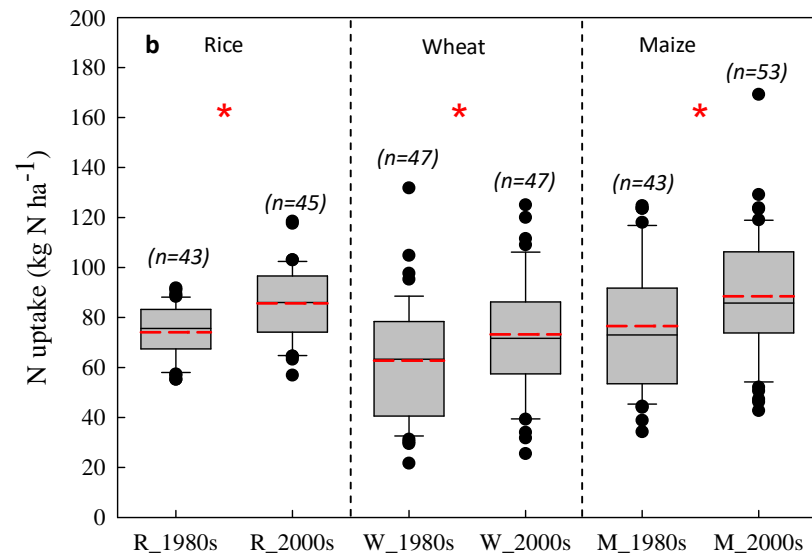
Increase: 5.9 mg/g (+30\%)

Crop N from zero-N plots

1980s: $71.1 \text{ kg N ha}^{-1}$

2000s: $82.4 \text{ kg N ha}^{-1}$

Increase: $11.3 \text{ kg N ha}^{-1 (+15\%)}$



These results suggest N enrichment in China's terrestrial ecosystems since 1980.

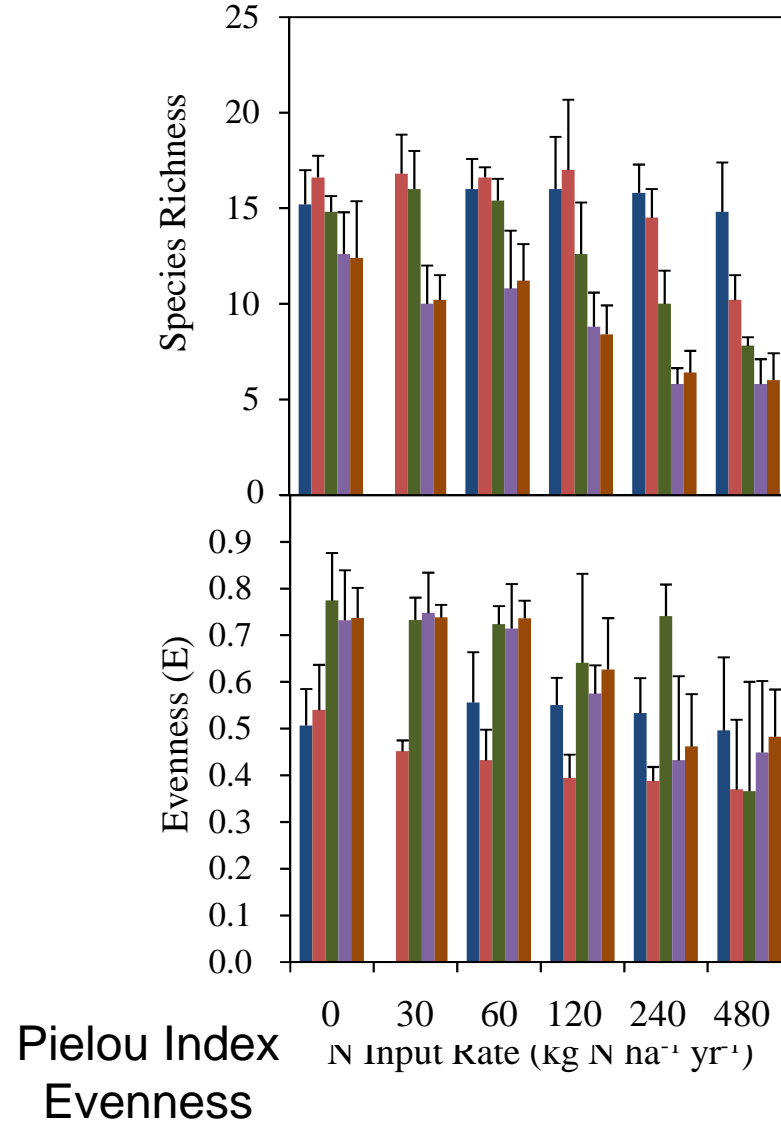
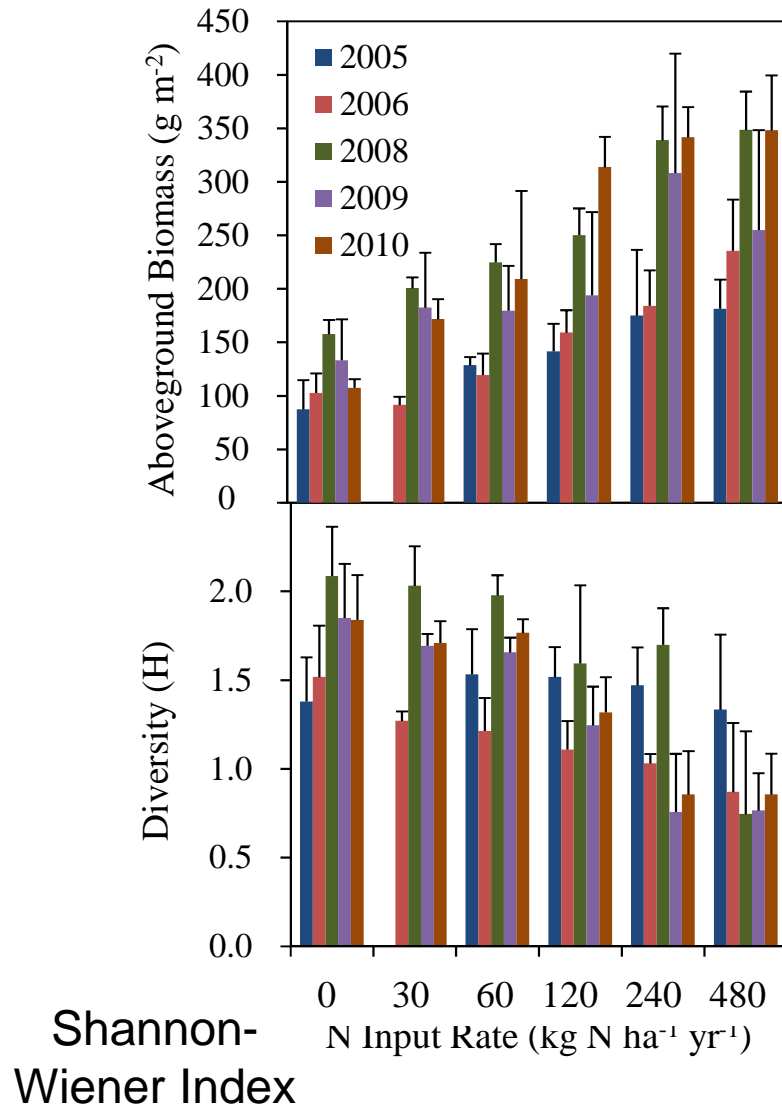
*, *** denote significant difference at 0.05, 0.001 level.

(Liu et al., 2013. Nature, 494:459-462)

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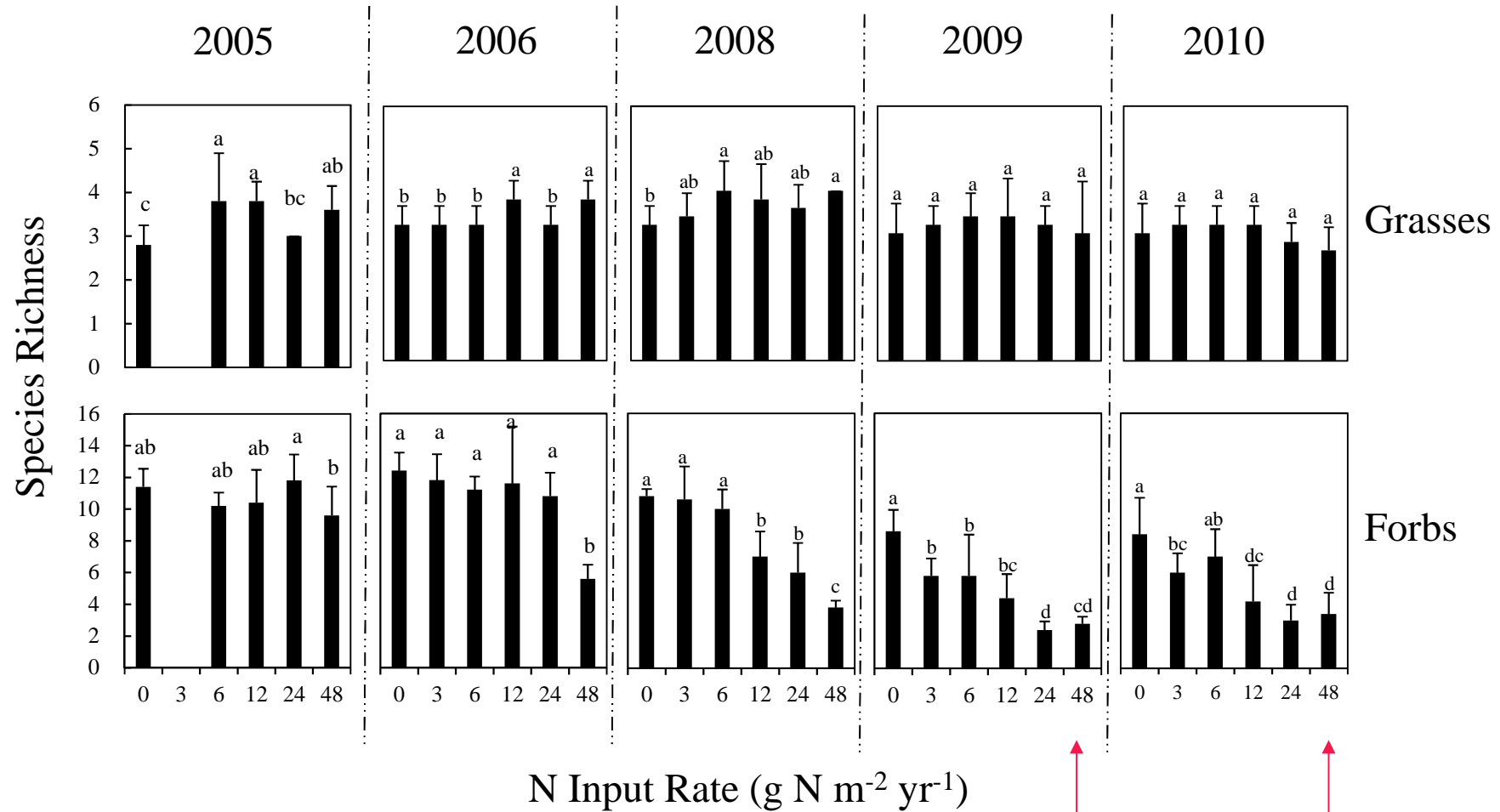
Impact of N addition on grassland plant community



(Song et al, 2012. J. Arid Land)

Elevated N deposition (e.g. 30 kg N ha⁻¹) led to significant decline in forb species richness in temperate grassland (Duolun, IM)

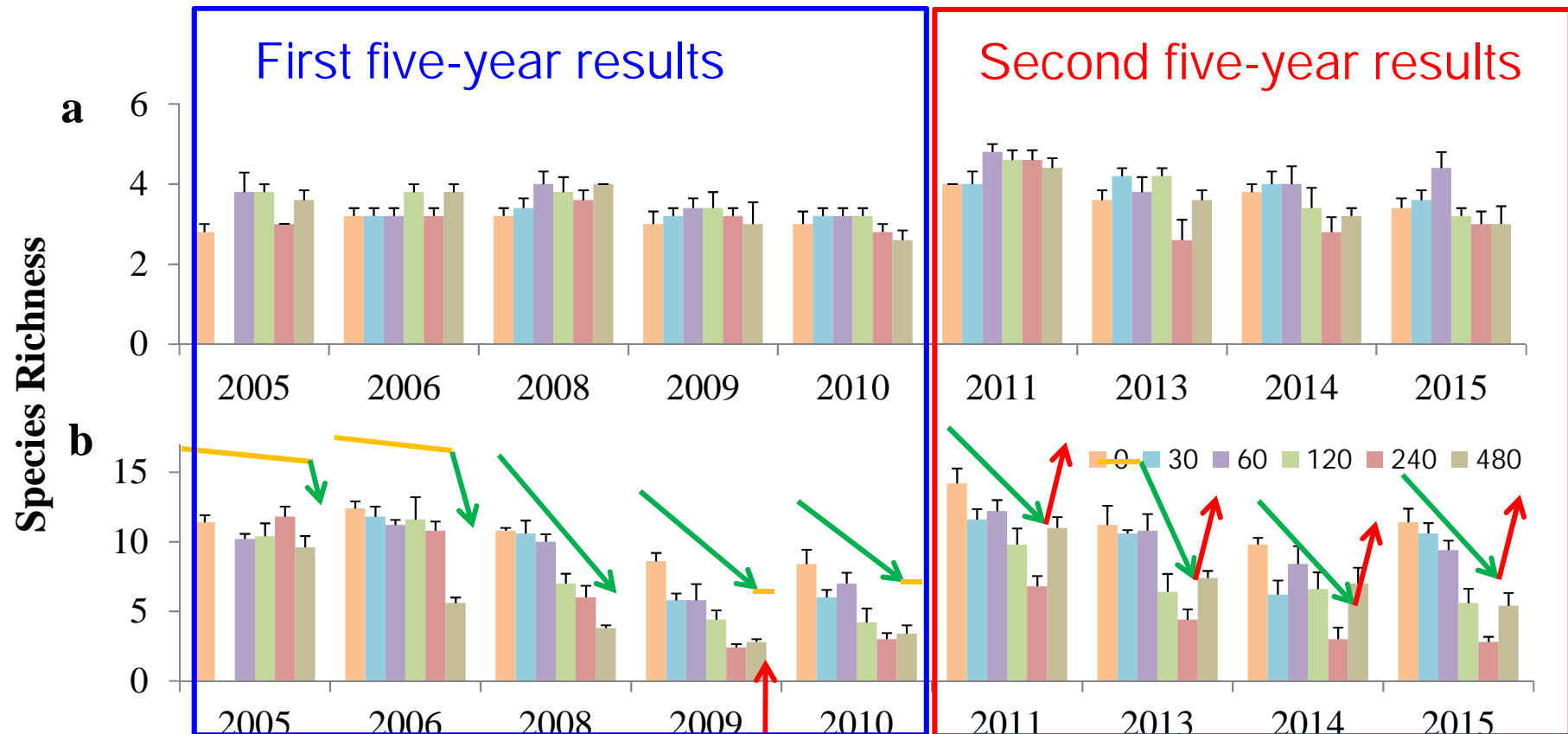
Relationship between N addition rate and species richness



Cessation of N addition

(Song et al, 2011, Biogeosciences)

Biodiversity loss and recovery under various N addition conditions



Relationship between N addition rate and species richness for grasses (a) and forbs (b) from 2005 to 2015.

(Hao et al., 2016. unpublished)

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Significant Acidification in Major Chinese Croplands

J. H. Guo, *et al.*

Science 327, 1008 (2010);

DOI: 10.1126/science.1182570

- Average pH decline was 0.5 units during 1980s and 2000s
- Soil pH decline: Cash crop systems > Cereal crop systems

Table 1. Topsoil pH changes in major Chinese croplands between the 1980s and 2000s. The soil groups are defined in (13). NS, not significant; pH range is an average (5 to 95 percentile).

Soil group	1980s		2000s					
	Sample number	pH value	Cereal crop systems*			Cash crop systems†		
			Sample number	pH value	pH change	Sample number	pH value	pH change
I	301	5.37 (4.40–6.60)	505	5.14 (4.17–6.52)	-0.23‡	337	5.07 (3.93–6.44)	-0.30‡
II	1157	6.33 (5.00–8.04)	1101	6.20 (5.00–7.70)	-0.13‡	413	5.98 (4.58–7.49)	-0.35‡
III	297	6.42 (4.50–8.30)	211	5.66 (4.27–8.06)	-0.76‡	98	5.62 (4.27–7.73)	-0.80‡
IV	562	6.32 (5.10–7.89)	537	6.00 (4.84–7.60)	-0.32‡	238	5.60 (4.07–7.42)	-0.72‡
V	995	7.96 (6.39–8.80)	850	7.69 (5.37–8.70)	-0.27‡	520	7.38 (5.69–8.20)	-0.58‡
VI	493	8.16 (7.10–8.80)	250	8.16 (7.49–8.82)	-0.00 (ns)	10	8.17 (7.43–8.93)	0.01 (ns)

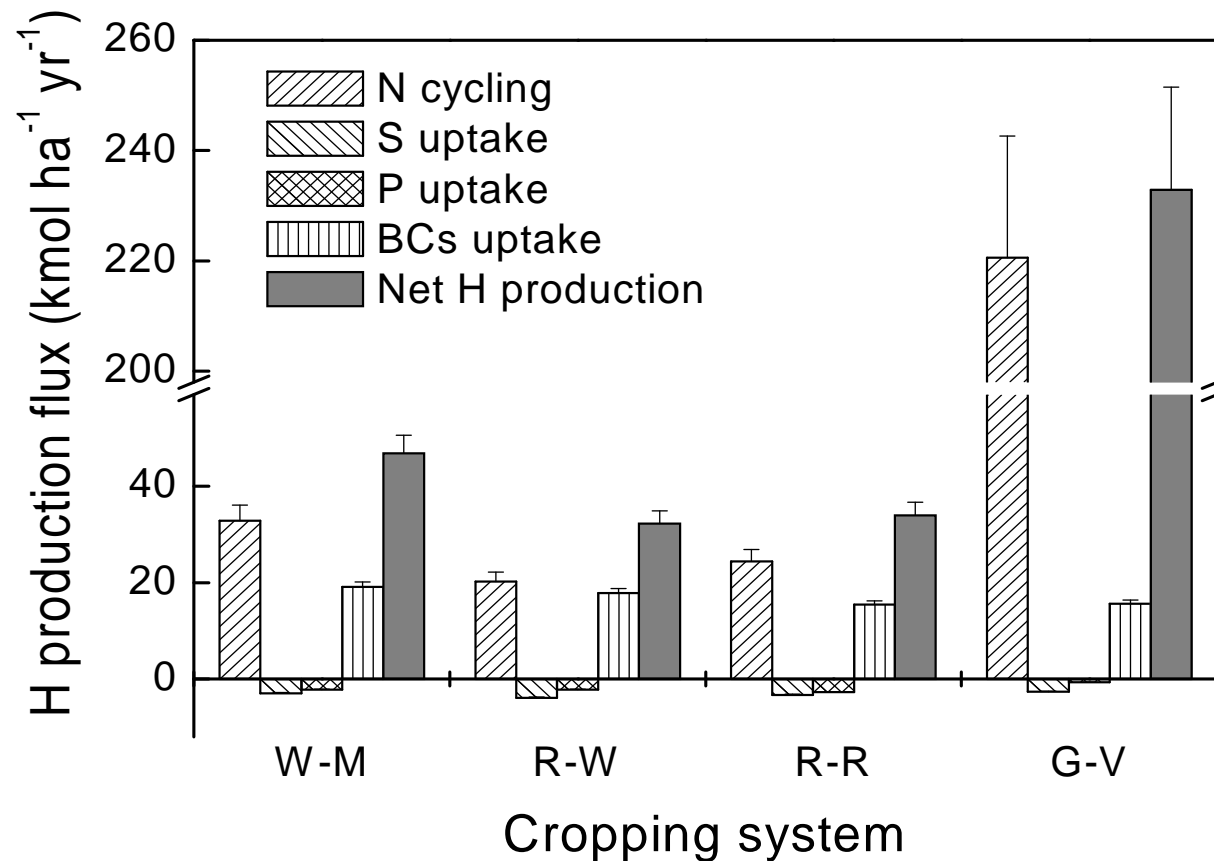
*Cereal/fiber crops (such as rice, wheat, maize, and cotton).

†High-input cash crops (such as vegetables, fruit trees, and tea).

‡P < 0.001.

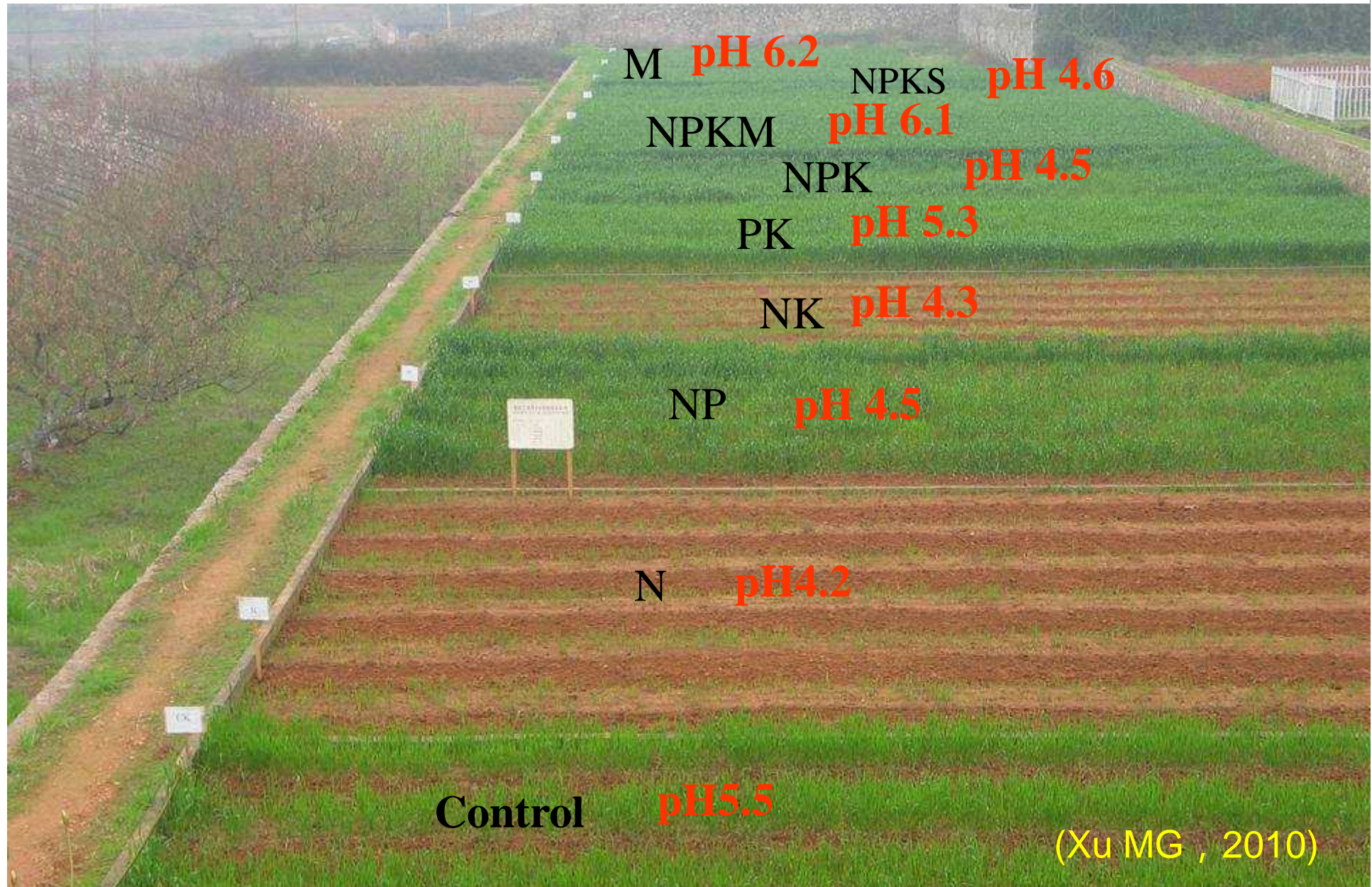
(Guo, Liu, Zhang, et al., 2010. *Science* 327: 1008-1010)

H⁺ production budget of main factors in four typical Chinese cropping systems. W-M: Wheat–maize; R-W: Rice-wheat; R-R: Rice-rice; G-V: Greenhouse vegetables.



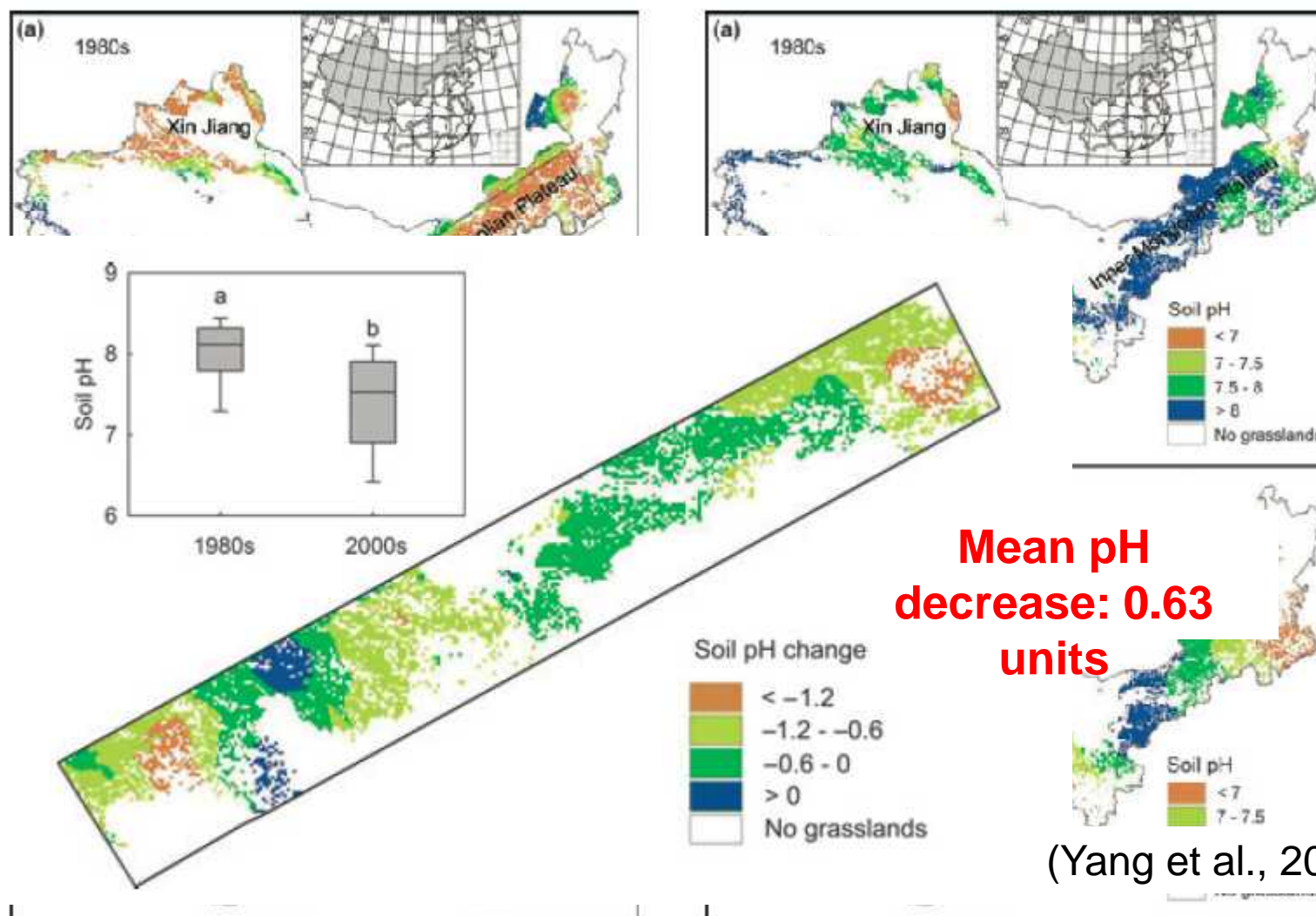
(Guo, Liu, Zhang et al., 2010. Science, 327: 1008-1010)

Changes of soil pH after 12-year fertilization in a wheat-maize cropping system (Annual N input 300 kg N/ha, Initial pH 5.7 in 1990)

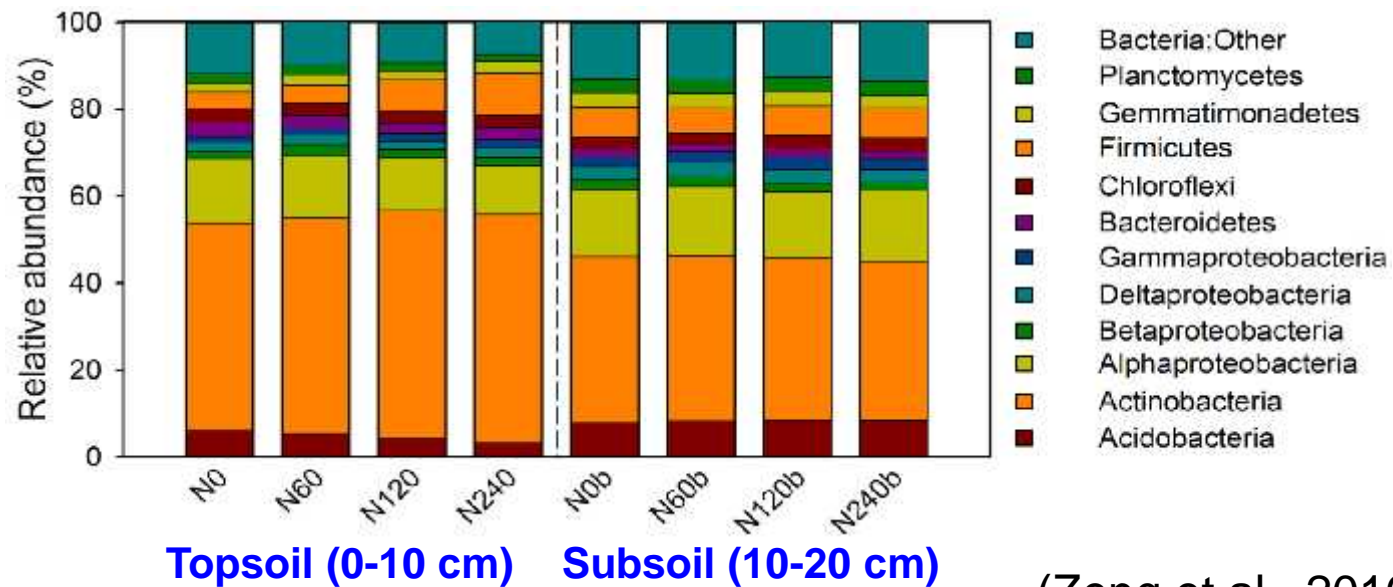
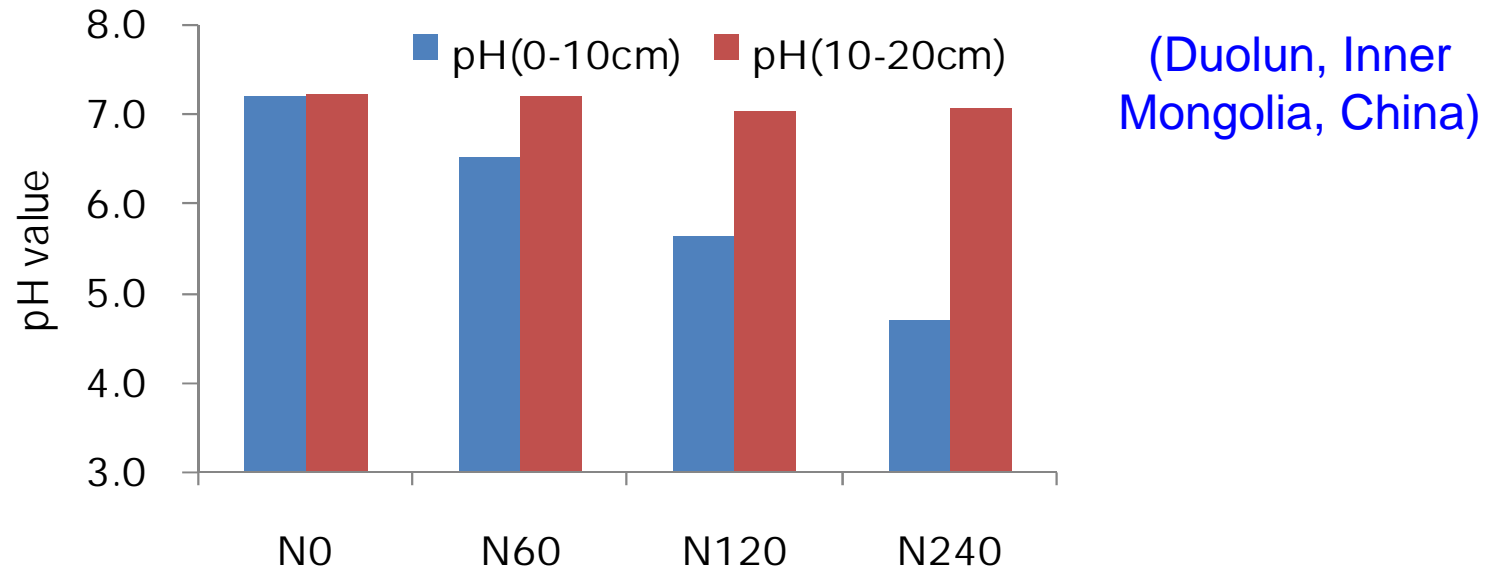


Significant soil acidification across northern China's grasslands during 1980s–2000s

YUANHE YANG*†‡, CHENGJUN JI*, WENHONG MA§, SHIFENG WANG‡, SHAOPENG WANG*, WENXUAN HAN¶, ANWAR MOHAMMAT||, DAVID ROBINSON‡ and PETE SMITH‡



Continuous N addition led to significant soil acidification and changes in microorganism communities

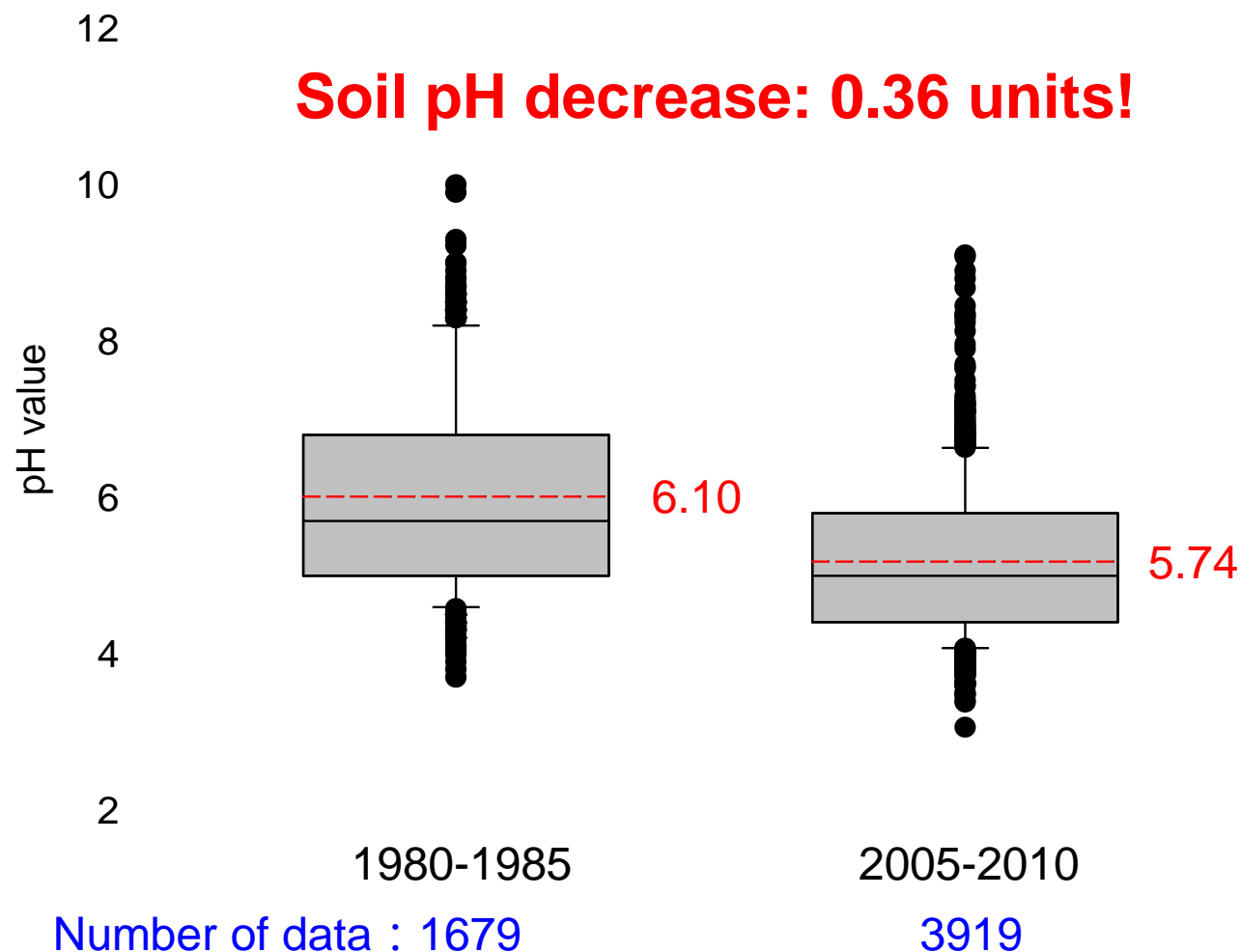


(Zeng et al., 2016. SBB)



The contribution of atmospheric deposition and forest harvesting to forest soil acidification in China since 1980

Qichao Zhu ^{a,1}, Wim De Vries ^{b,c}, Xuejun Liu ^{d,e}, Mufan Zeng ^a, Tianxiang Hao ^a, Enzai Du ^f, Fusuo Zhang ^g, Jianbo Shen ^h



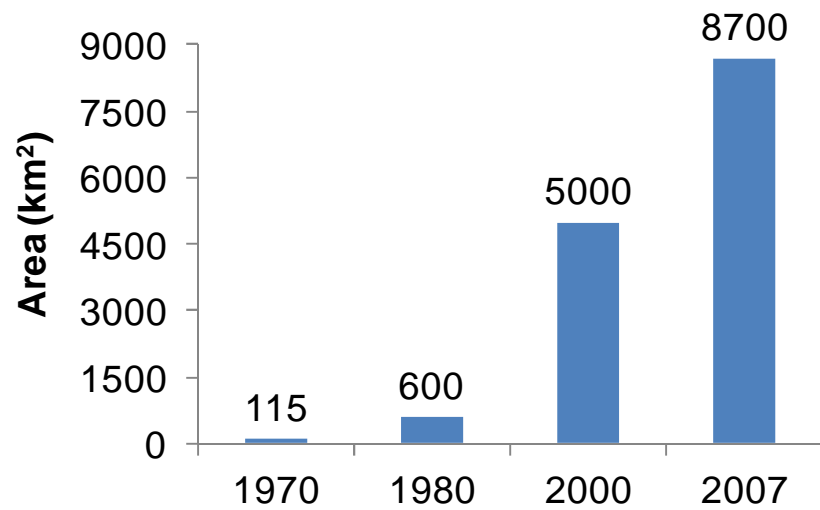
Atmospheric deposition contributes to 84% of H⁺ production in forest soils.

(Zhu et al. 2016. AE)

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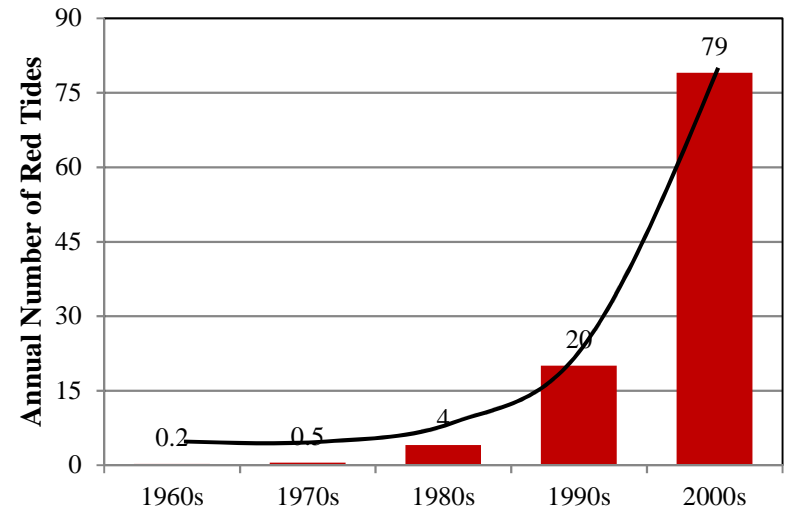
China's eutrophied lake area has increased from 135 (1970) to 8700 km² (2007).



(Jin, 2009)



太湖蓝藻暴发 2007年11月2日11时00分 (新华社)



(the State Oceanic Administration, 2009)

(*Science* 2009, 1014-1015)

POLICYFORUM

ECOLOGY

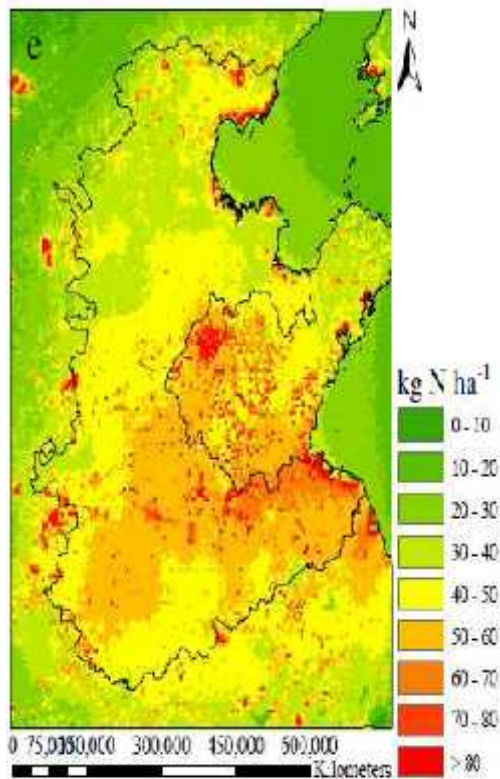
Controlling Eutrophication: Nitrogen and Phosphorus

Daniel J. Conley,^{1*} Hans W. Paerl,² Robert W. Howarth,³ Donald F. Boesch,⁴ Sybil P. Seitzinger,⁵ Karl E. Havens,⁶ Christiane Lancelot,⁷ Gene E. Likens⁸

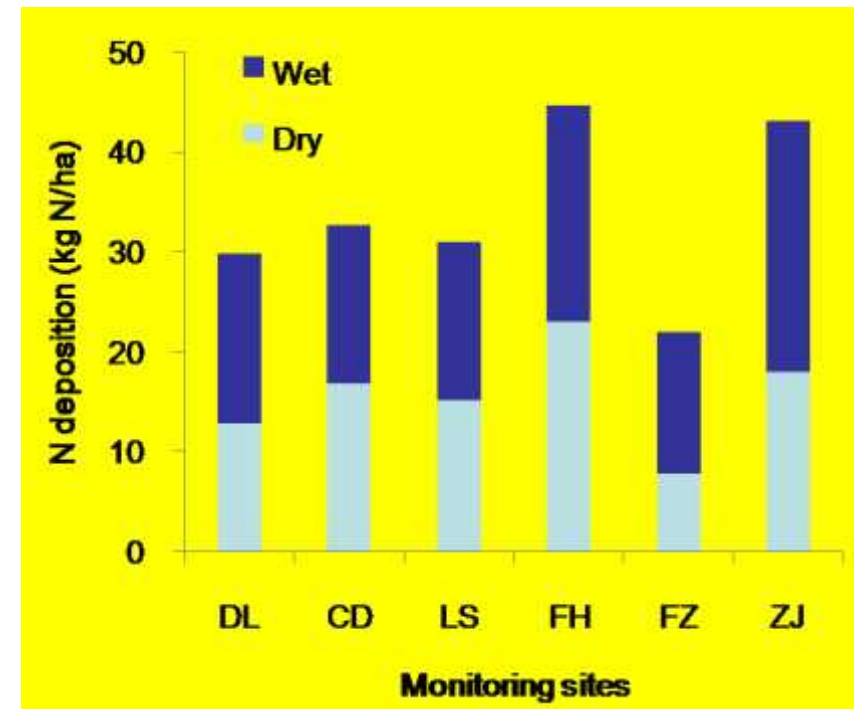


N deposition onto coastal seas of China

Total N deposition to Bohai Sea: 22 kg N ha⁻¹ yr⁻¹



Annual N deposition in coastal zone: 34 kg N ha⁻¹



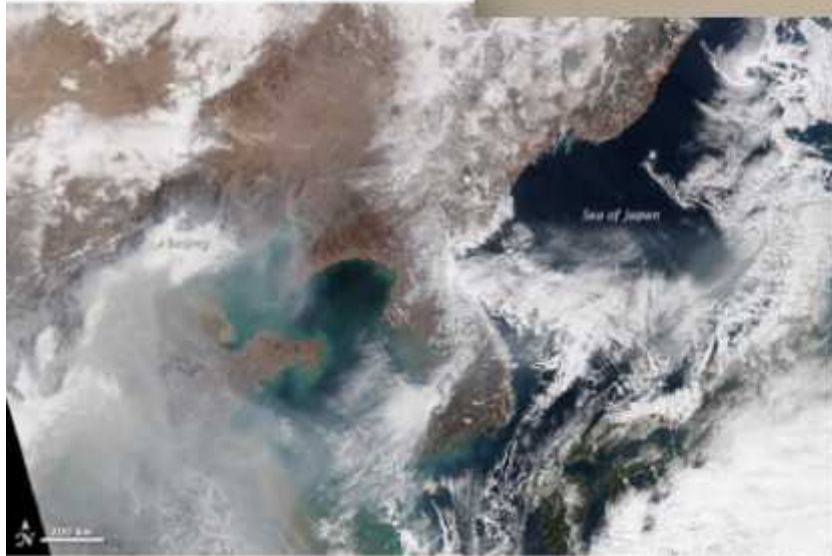
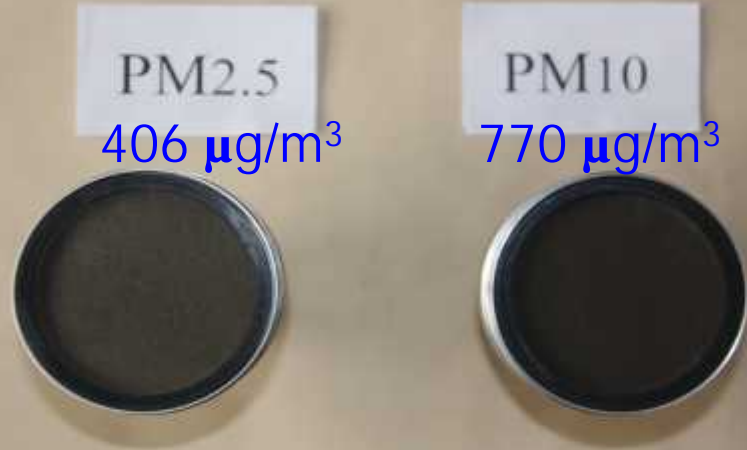
Assuming half of coastal N deposition rate ($\bullet 17$ kg N/ha) onto Bohai, Yellow, East China and South China Seas (3×10^6 km²), total N deposition amounts to 5 Tg N yr⁻¹ to the 4 marine ecosystems surrounding China.

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Beijing Smog 2014

CAU west campus, Feb. 26th, 2014



acquired February 26, 2014



Enhanced nitrogen deposition over China

Xuejun Liu^{1*}, Ying Zhang^{1*}, Wenxuan Han¹, Aohan Tang¹, Jianlin Shen¹, Zhenling Cui¹, Peter Vitousek², Jan Willem Frisman^{3,4}, Keith Goulding⁵, Peter Christie^{1,6}, Andreas Fangmeier⁷ & Fusuo Zhang¹

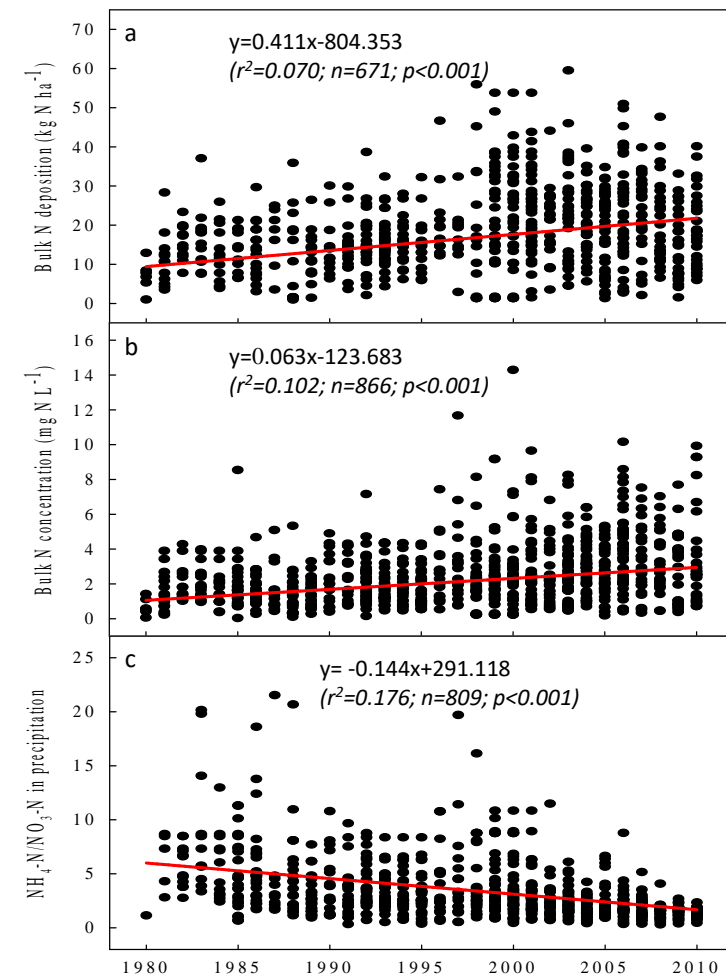
1980s: 13.2 kg N ha⁻¹

2000s: 21.1 kg N ha⁻¹

Increase: • 8 kg N ha⁻¹ or 60%



(Liu and Zhang et al., 2013. Nature 294: 259-262)

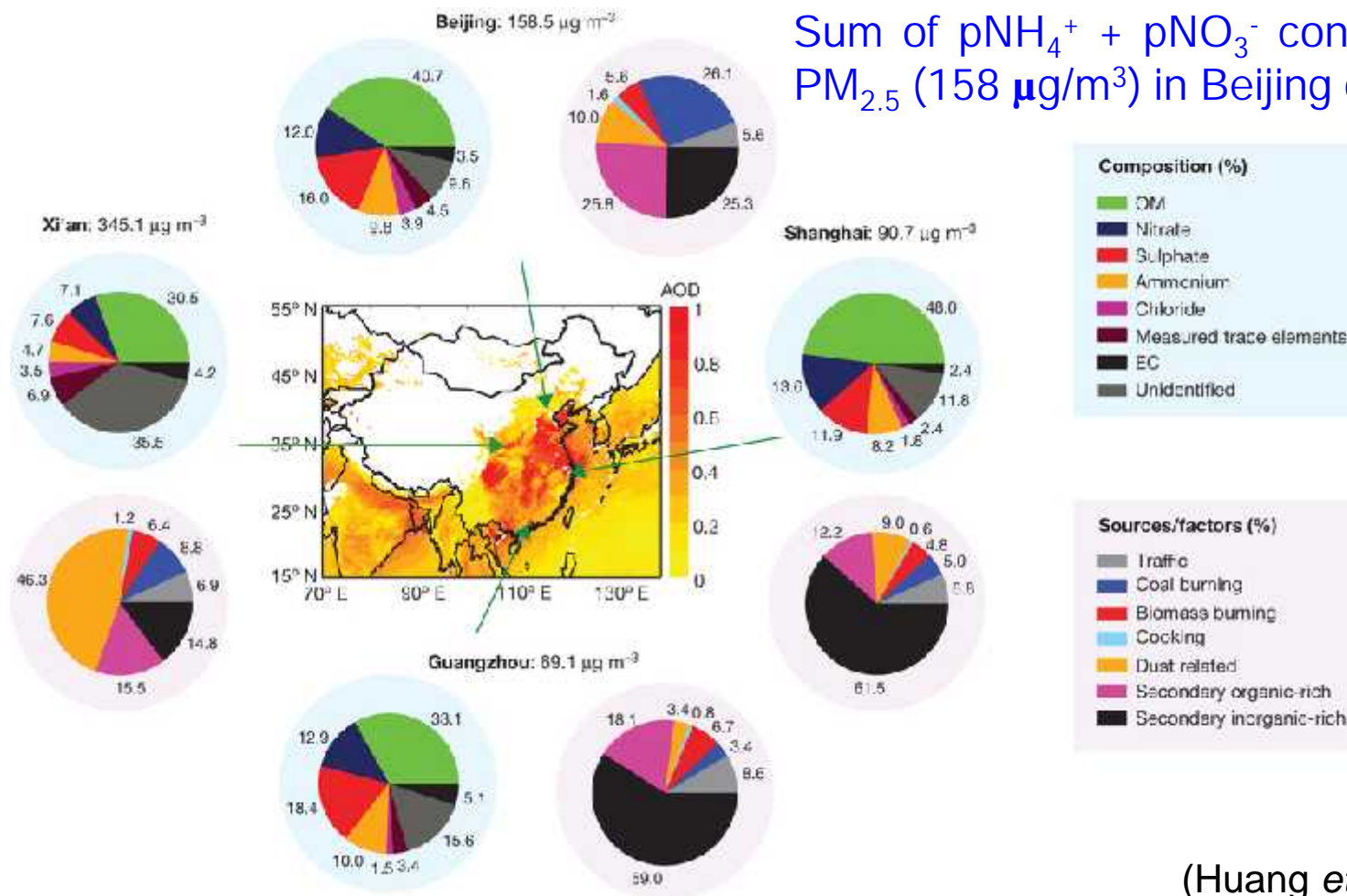


High secondary aerosol contribution to particulate pollution during haze events in China

Ru-Jin Huang^{1,2*}, Yanlin Zhang^{3,4}, Carlo Bozzetti¹, Kin-Fai Ho⁵, Jun-Ji Cao², Yongming Han², Kaspar R. Daellenbach¹, Jay G. Slowik¹, Stephen M. Platt¹, Francesco Canonaco¹, Peter Zotter¹, Robert Wolf¹, Simone M. Dieber¹, Emily A. Brunst¹, Monica Crippa^{6†}, Giancarlo Ciarelli⁴, Andrea Piazzalunga⁴, Margit Schwikowski^{1,7}, Güleln Abbaspour⁸, Jürgen Schnell-Kretz⁹, Ralf Zimmermann¹⁰, Zhisheng An¹¹, Sönke Szidat¹, Urs Baltensperger¹, Imad El Haddad^{1*} & André S. H. Prévôt¹

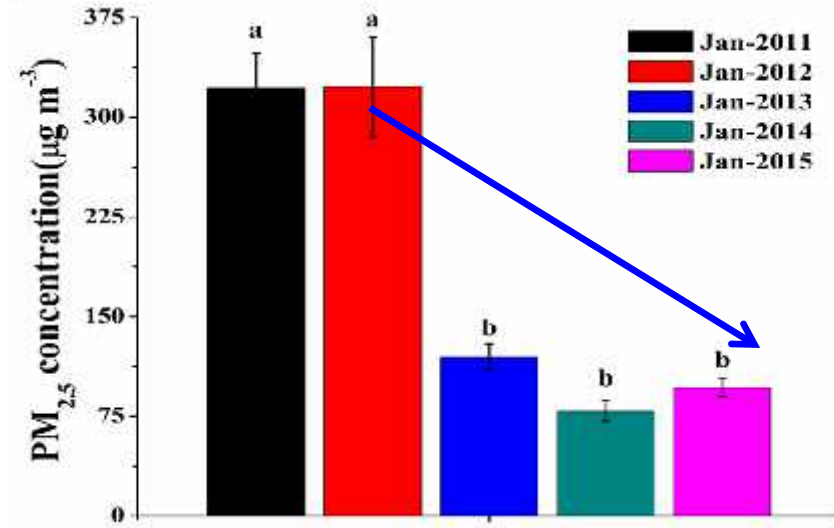
Chemical composition and source apportionment of PM_{2.5} during the high pollution events of 5–25 January 2013 in Beijing, Shanghai, Guangzhou and Xi'an cities

Sum of pNH₄⁺ + pNO₃⁻ contributed to 23% of PM_{2.5} (158 μg/m³) in Beijing during Jan. 2013.

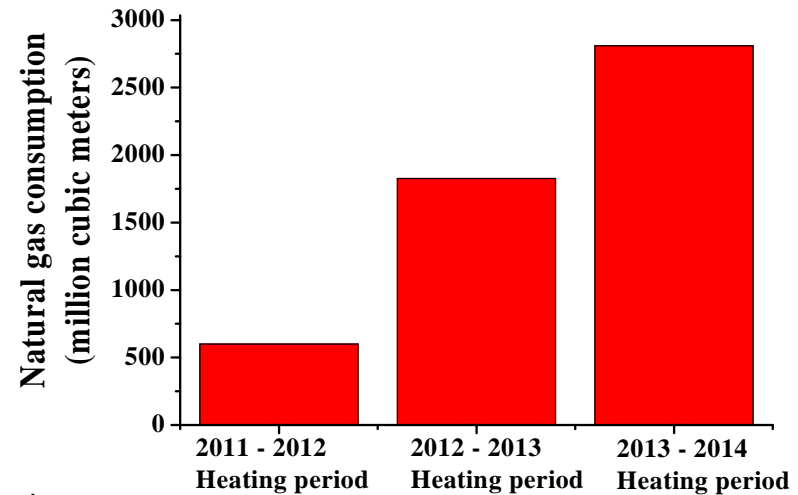
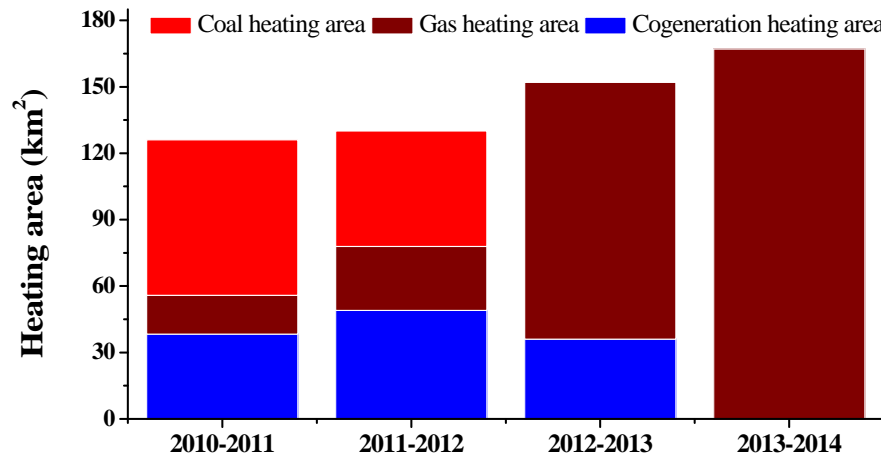


(Huang *et al.*, 2014. Nature)

A successful example at Urumqi: Shift of coal to natural gas for heating



N dry deposition decreased from 26 kg N ha⁻¹ (2012) to 16 kg N ha⁻¹ (2015).



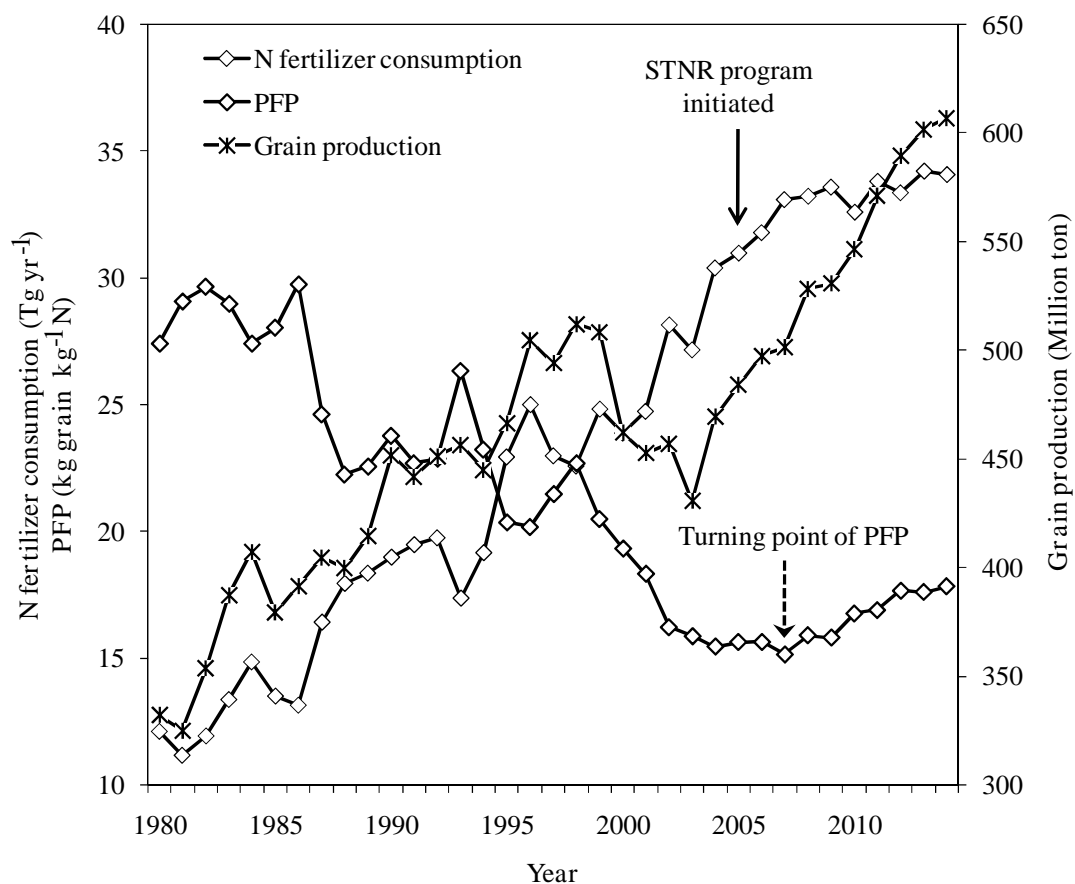
(Song et al., 2015. EST 49: 2066-2072)

Conclusions and outlook

- Overall N input/output ratio is 43% currently, with an slight increasing trend in recent years;
- Huge regional differences in the trend exist due to various agricultural structure and soil conditions;
- Enhanced N cycling has impacted largely to both agricultural and natural ecosystems, including biodiversity loss, soil acidification, eutrophication, air pollution, etc.;
- Knowledge based practices in improving N use are promising, although barriers still exist.

Evidence for a Historic Change Occurring in China

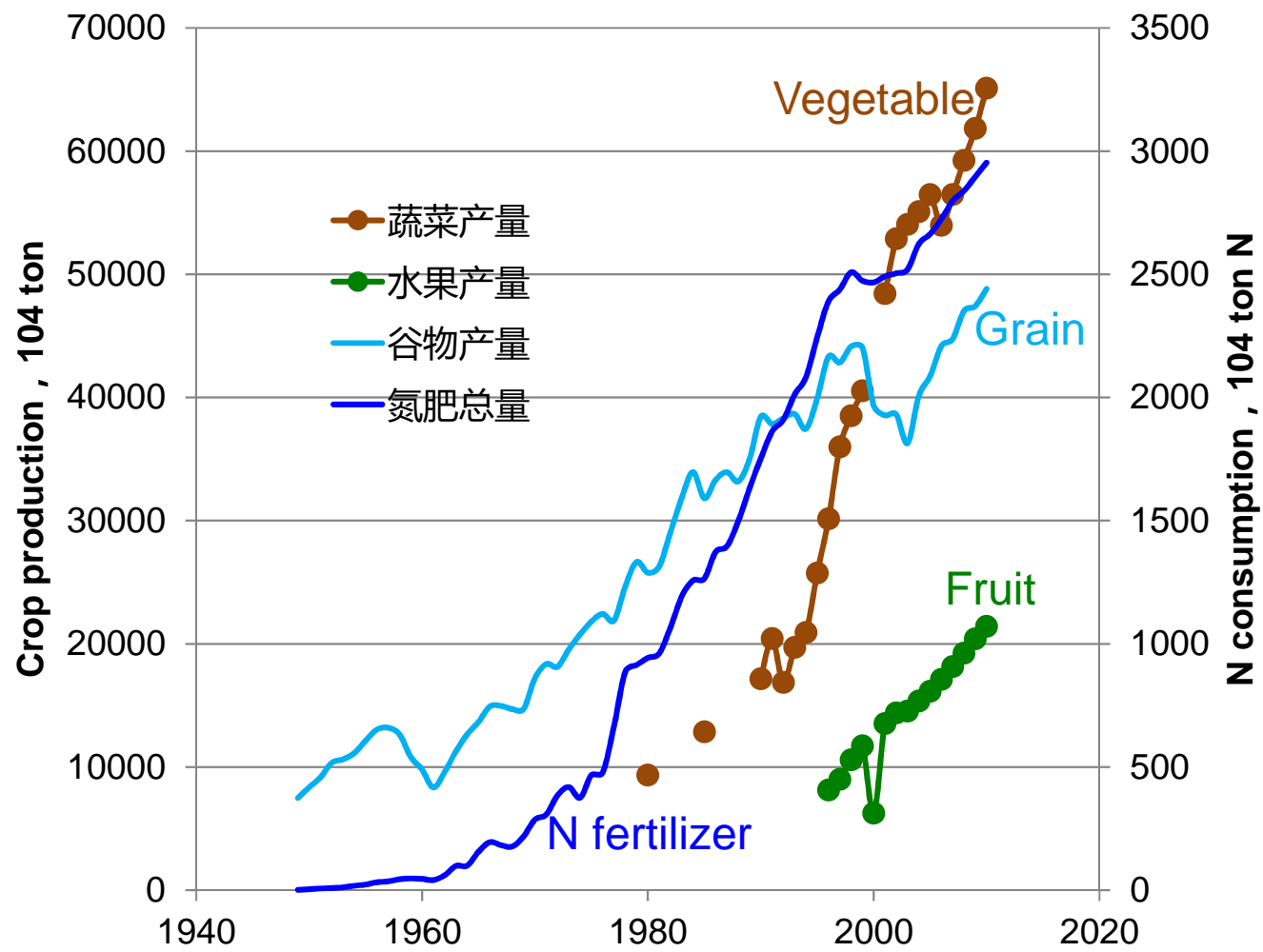
Xuejun Liu,^{*,†} Peter Vitousek,[‡] Yunhua Chang,[§] Weifeng Zhang,[†] Pamela Matson,^{||} and Fusuo Zhang[†]



China is now at a historic turning point to increase NUE and crop production meanwhile reducing its N environmental impact.

(Liu et al., 2016. EST)

**Thanks for
your attention!**



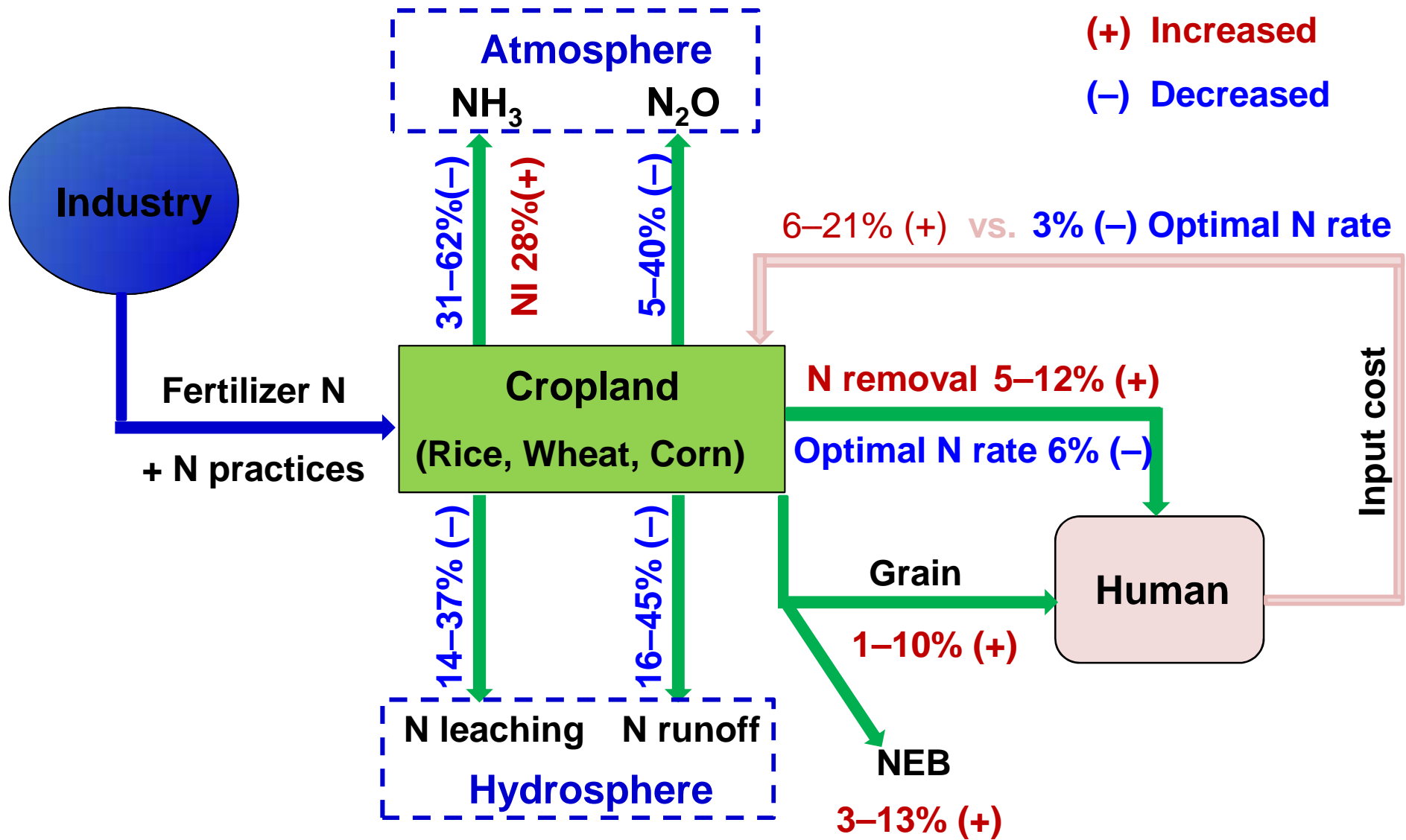
Implementation

County Level Arable Land Resource Management Information System

Developed by
Yangzhou Soil and Fertilizer Station



Overall effects of N practices



Nitrogen input to croplands

